

Addressing temporal considerations in life cycle assessment

Didier Beloin-Saint-Pierre, Ariane Albers, Arnaud Hélias, Ligia Tiruta-Barna, Peter Fantke, Annie Levasseur, Enrico Benetto, Anthony Benoist, Pierre Collet

▶ To cite this version:

Didier Beloin-Saint-Pierre, Ariane Albers, Arnaud Hélias, Ligia Tiruta-Barna, Peter Fantke, et al.. Addressing temporal considerations in life cycle assessment. Science of the Total Environment, 2020, pp.140700. 10.1016/j.scitotenv.2020.140700. hal-02894987v1

HAL Id: hal-02894987 https://ifp.hal.science/hal-02894987v1

Submitted on 9 Jul 2020 (v1), last revised 15 Sep 2020 (v2)

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 Addressing temporal considerations in life cycle assessment

- 2 Didier Beloin-Saint-Pierre^{1†}, Ariane Albers², Arnaud Hélias³, Ligia Tiruta-Barna⁴, Peter Fantke⁵,
- 3 Annie Levasseur ⁶, Enrico Benetto ⁷, Anthony Benoist ⁸, Pierre Collet ²
- 4 † Corresponding author
- 5 ¹ Empa Materials Science and Technology, Lerchenfeldstrasse 5, CH-9014 St. Gallen, Switzerland
- 6 ² IFP Energies Nouvelles, 1 et 4 Avenue de Bois-Préau, 92852 Rueil-Malmaison, France
- 7 ³ ITAP, Irstea, Montpellier SupAgro, Univ Montpellier, ELSA Research Group, Montpellier, France
- 8 ⁴ TBI, Université de Toulouse, CNRS, INRAE, INSA, Toulouse, France
- ⁵ Quantitative Sustainability Assessment, Department of Technology, Management and Economics, Technical
 University of Denmark, Kgs. Lyngby, Denmark
- ⁶ École de technologie supérieure, Construction Engineering Department, 1100 Notre-Dame West, Montréal,
 Québec, Canada
- ⁷ Environmental Sustainability Assessment and Circularity Unit, Department of Environmental Research and
 Innovation, Luxembourg Institute of Science and Technology, Esch/Alzette, Luxembourg
- 15 ⁸ CIRAD, UPR BioWooEB, F-34398 Montpellier, France

16 HIGHLIGHTS:

- 17 Review of temporal considerations in the life cycle assessment methodology
- Glossary of important terms for time considerations in life cycle assessment
- Key aspects of dynamic life cycle assessments
- 20 Current implementation challenges for dynamic life cycle assessment
- 21 Development pathways for future dynamic life cycle assessment
- 22

Technosphere	Dynamic of systems			
Dynamic LCI	Variations	Evolution		
	Occurrence	Temporal scope		
Dynamic LCA	Differentiation	Prospective modelling		
Dynamic LCIA	Resolution	Period of validity		
Ecosphere	Time	horizon		

24

25 ABSTRACT

26 In life cycle assessment (LCA), temporal considerations are usually lost during the life cycle inventory 27 calculation, resulting in an aggregated "snapshot" of potential impacts. Disregarding such temporal 28 considerations has previously been underlined as an important source of uncertainty, but a growing number of 29 approaches have been developed to tackle this issue. Nevertheless, their adoption by LCA practitioners is still 30 uncommon, which raises concerns about the representativeness of current LCA results. Furthermore, a lack of 31 consistency can be observed in the used terms for discussions on temporal considerations. The purpose of this 32 review is thus to search for common ground and to identify the current implementation challenges while also 33 proposing development pathways.

This paper introduces a glossary of the most frequently used terms related to temporal considerations in LCA to build a common understanding of key concepts and to facilitate discussions. A review is also performed on current solutions for temporal considerations in different LCA phases (goal and scope definition, life cycle inventory analysis and life cycle impact assessment), analysing each temporal consideration for its relevant conceptual developments in LCA and its level of operationalisation.

We then present a potential stepwise approach and development pathways to address the current challenges of implementation for dynamic LCA (DLCA). Three key focal areas for integrating temporal considerations within the LCA framework are discussed: i) define the temporal scope over which temporal distributions of emissions are occurring, ii) use calendar-specific information to model systems and associated impacts, and iii) select the appropriate level of temporal resolution to describe the variations of flows and characterisation factors.

Addressing more temporal considerations within a DLCA framework is expected to reduce uncertainties and increase the representativeness of results, but possible trade-offs between additional data collection efforts and

46 the increased value of results from DLCAs should be kept in mind.

47 KEYWORDS:

48 Dynamic LCA, temporal considerations, review, recommendations, implementation challenges

- 50
- 51

52 1. INTRODUCTION

53 Disregarding temporal considerations¹ has been identified as an inherent limitations of life cycle assessment 54 (LCA) (ISO14040, 2006; ISO14044, 2006). Indeed, the importance of properly considering the dynamics of 55 environmental sustainability for the comparison of products, services or systems has been explored, debated and 56 confirmed during the last 20 years by many researchers like Owens (1997a), Herrchen (1998), Reap et al. 57 (2008a; 2008b), Finnveden et al. (2009), Levasseur et al. (2010) and McManus & Taylor (2015), to name a few. 58 In this discussion, Rebitzer et al. (2004), Reap et al. (2008a) and Yuan et al. (2015) have mainly explored the 59 subject of dynamics in human activities. During the same period, Reap et al. (2008b), Shah & Ries (2009), 60 Fantke et al. (2012), Kendall (2012), Levasseur et al. (2012b) and Manneh et al. (2012) have proposed different 61 ideas on the dynamics of environmental responses to human pressures. Additionally, Hellweg et al. (2003b; 62 2005; 2014), Levasseur et al. (2013), Saez de Bikuña et al. (2018) and Yu et al. (2018) have underlined different 63 potential effects from the choice of temporal boundaries in LCA studies. These three general subjects have 64 covered the bulk of the conversation on temporal considerations in the LCA framework and a growing awareness of the LCA community on this topic is shown in figure 1^2 with a growth in the number of publications 65 66 where some aspects are addressed.



¹ Consideration encompass all aspects relating to the description of time and dynamics of systems (see glossary in Table 1).

² The annual number of publications were found with the advance search function on web of science. The following words and conditions were searched for in the topic section: ("life cycle assessment" AND temporal) + ("life cycle assessment" AND "time horizon") + ("life cycle assessment" AND dynamic). The word "time" was not part of the search to avoid mentions of the time required for data gathering activity and because it can be part of words like "sometimes". The search was made on the 17 of December, 2019.

Figure 1: Numbers of LCA publications per year where temporal considerations are discussed

67 Within the identified 1281 publications, 53 review papers present several discussions about temporal 68 considerations in different sectors (e.g. agriculture, building and energy) or in the general LCA framework. Very 69 recently, Sohn et al. (2020) and Lueddeckens et al. (2020) have proposed reviews on aspects or issues that are 70 connected to the approach of dynamic LCA (DLCA). In Sohn et al. (2020), three types of dynamism have been 71 defined: dynamic process inventory, dynamic system inventory and dynamic characterisation, thus focusing on 72 the concern of changes in human activities and environmental responses with many implementation examples. 73 Lueddeckens et al. (2020) have offered a clearly structured analysis of 60 documents that have been published 74 until the end of 2018 where interdependencies are underlined and solutions from the literature are identified for 75 six types of temporal issues (i.e. time horizon, temporal weighting/discounting, temporal resolution of the 76 inventory, time-dependent characterisation, dynamic weighting and time-dependent normalisation). While 77 comprehensive for these six issues, the work of Lueddeckens et al. (2020) does not offer a detailed discussion on 78 questions like computation, uncertainty and variability for the DLCA approach.

79 When looking at the abundant literature on the subject of temporal considerations in LCA, it rapidly becomes 80 clear that the vocabulary in recent and older reviews varies considerably for common aspects such as the 81 temporal scope or time horizon. We believe that this lack of consistency in terminology is hindering a clear 82 discussion on the subject and therefore the development of new propositions that can be accepted by a majority 83 of researchers. Furthermore, while many ideas, concepts, approaches and tools have been suggested by 84 researchers and are now used in publications under the term DLCA, their widespread implementation by 85 practitioners is still far from reached. This lack of temporal considerations in most LCA studies is worrisome 86 since it was shown that such aspects may have significant effects on LCA results mainly in the sectors of 87 buildings (Collinge et al., 2018; Negishi et al., 2019; Roux et al., 2016b) and energy (Amor et al., 2014; Beloin-88 Saint-Pierre et al., 2017; Menten et al., 2015; Pehnt, 2006). It thus seems important to identify and address the 89 current implementation challenges that prevent LCA practitioners from more frequent accounting of temporal 90 considerations.

These challenges are tackled in the following sections. First, a glossary in section 2 proposes definitions for terms related to temporal considerations in LCA, which should clarify shared aspects of past discussions and help in building consensus. These terms are then used consistently in the text. Section 3 follows with a review of the LCA literature that highlights current implementation challenges for a broad application of the DLCA

- 95 approach. Recommendations for current implementation options and further developments are then provided in
- section 4.
- 97

98 Finding a clear structure to organise and analyse the numerous options for temporal consideration that have been 99 discussed in the last 20 years of LCA development can be a daunting task. Previous reviews have chosen 100 different strategies mainly based on specific sectors, themes or issues. These scheme have often limited the scope of the analysis or the identification of connections between ideas. We therefore chose another perspective 101 102 that classifies temporal considerations based on why they are used (i.e. purposes). Indeed, from our 103 understanding, temporal considerations are employed in LCA studies to define the temporal scope, to describe 104 the dynamic of systems and to increase the representativeness of models. We also differentiate the temporal 105 considerations within the standard phases of the LCA framework to provide a frame of reference that is well-106 known to practitioners. We thus hope to cover most options for temporal consideration in LCA with this strategy 107 and to comprehensively address the topic for a broader implementation of DLCA studies in the future.

108 2. PROPOSED GLOSSARY

Table 1 proposes key terms and definitions to discuss temporal considerations within the LCA framework. These terms are used throughout this review to ensure a consistent and non-ambiguous discussion for future developments. It is also the authors' hope that this glossary might bring some uniformity in future discussions. Concepts behind the most recently proposed definitions for types of dynamism and four subtypes of DLCA (Sohn et al., 2020) can be found in this table with a somewhat different perspective.

115 Table 1: List of proposed terms defining key temporal considerations in the LCA framework. The list is in alphabetical order

	116	so all terms from	this glossary	v are underlined to	highlight the links.	Words in	brackets are syn	onyms from the literatur
--	-----	-------------------	---------------	---------------------	----------------------	----------	------------------	--------------------------

Term	Definition
Dynamic LCA	LCA studies where relevant dynamic of systems and/or temporal differentiation of flows
(DLCA)	are explicitly defined and considered.
Dynamic LCI	Life cycle inventory (LCI) that is calculated from supply and value chains where dynamic
(DLCI)	of systems or temporal differentiation is considered, resulting in temporal distributions to
	describe elementary flows.
Dynamic LCIA	Characterisation models of environmental mechanisms that account for the dynamic of
(DLCIA)	ecosphere systems and can therefore use temporal information of DLCIs. The chosen
	temporal differentiation (e.g. day, season, and year) can depend on the impact categories.
	Both case specific and calendar-based characterisation models can be used, depending on
	the chosen indicators.
Dynamic of systems	System modelling that considers inherent variations, periods of occurrence or evolution
	within the temporal scope of models' components. Such a dynamic modelling can be
	applied to both technosphere systems (for LCI) and ecosphere systems (for LCIA).
Evolution	Changes of process, structure or state models' components (e.g. technology replacement,
	pollutant concentration in a compartment of the environment).
Inherent variations	Variations of flows in the models' components (e.g. cycles of solar energy production,
	growth rates of vegetation, seasonal functional traits, biogeochemical and biophysical
	dynamics). The discontinuities of flow rates are also part of such changes.

Models' components	Information structuring all models. At the technosphere level, components are elementary flows, product flows and processes. At the ecosphere level, components of LCIA models differ between impact categories. For example, components for freshwater ecotoxicity can be environmental fate, ecosystem exposure and ecotoxicological effects (Fantke et al., 2018).
Period of occurrence	The moment when a <u>model's component</u> is starting, modified or finishing over time. (e.g. lifespan of a building, beginning of waste management, start of a life cycle)
Period-specific	CF for a given temporal scope or period of occurrence. It results from the dynamic of
characterisation factor (CF)	<u>systems</u> in the ecosphere and can be calendar-specific, relative to the length of the <u>temporal</u> <u>scope</u> , or defined by a <u>TH</u> . <u>Period-specific CFs</u> are modelled as constant over the chosen period.
Period of validity	The period over which datasets, LCIs or LCIA methods are considered valid representations. This information should be calendar-based. [Time context (ILCD), time frame, range of time, period of time, time period, timespan, temporal boundary, time scale and time horizon]
Prospective modelling	A prospective LCA addresses future life cycle impacts using different modelling strategies (e.g. scenario-based, technology development curves and agent- or activity-based models). The <u>evolution</u> of systems is thus defined and/or simulated using a list of explicit assumptions regarding the future. <u>Prospective modelling</u> can be applied to both the technosphere and ecosphere and is a subset of the <u>dynamic of systems</u> , which only concerns future forecasts.
Temporal considerations	Any aspects (i.e. information) described in relation to the time dimension or <u>dynamic of</u> <u>systems</u> in the LCA framework. This is the overarching term relating to all other terms of the glossary. [Time-aspect in ILCD documents]
Temporal differentiation	The action of distributing the information on a time scale related to the <u>models'</u> <u>components</u> . For example, elementary flows could be described per day or year. Different processes representing yearly average are another example. [Temporal segmentation in ILCD]
Temporal resolution	Describes the time granulometry when <u>temporal differentiation</u> is carried out. For instance, a monthly or daily resolution can be used to describe the flows in technosphere models. The same term can be used to describe a time step for <u>period-specific CFs</u> . [Time step]
Temporal	Qualitative or quantitative assessment of data, processes or LCIA methods in relation to
representativeness	how appropriate their information fits with their <u>temporal scope</u> . [Time-related representativeness (ILCD), Time-related coverage (ISO14044)]
Temporal scope	Defines any type of period that is considered in a LCA study (e.g. temporal considerations along a life cycle, service life of a product, data collection period).
Temporalisation	Attribution of temporal properties to the <u>models' components</u> . (e.g. definition of <u>temporal scopes</u>)
Time horizon (TH)	Relative <u>temporal scope</u> over which environmental impacts are summed up to provide LCA results.

117 3. TEMPORAL CONSIDERATIONS FOR DIFFERENT PURPOSES

118	Many temporal considerations have been described in previous publications, reports and standards to develop the
119	general LCA framework (ISO14040, 2006; ISO14044, 2006; Joint Research Center, 2010) and its dynamic
120	counterpart. For instance, Sohn et al. (2020) classified 56 DLCA studies by their technological domains and
121	types of assessed dynamism. In this section, the considerations are first regrouped by their purposes. A Venn
122	diagram in figure 2 presents this organisation of temporal considerations where gold, purple and red rounded
123	rectangles respectively highlight the purposes of defining the temporal scope, considering the dynamic of
124	systems and increasing the temporal representativeness. 10 classes of temporal considerations are also presented

- 125 with rectangles of different colours and linked to the phases of the LCA framework where they most commonly
- 126 appear. In figure 2, the interpretation phase is excluded because the identified temporal considerations are first
- 127 accounted for in the three mentioned phases and can then be used to analyse the results.



Figure 2: Venn diagram of temporal considerations in relation to their purposes (grey rectangles), the phases of the LCA methodology (coloured rectangles) and 10 classes (Bold titles). Existing connections are presented by arrows.

128

- The level of relevance, conceptual development and operationalisation for the temporal considerations of figure 2 are qualitatively assessed with scores ranging from A (highest) to C (lowest) (detailed in table 2) to evaluate the state-of-the-art shown in table 3. A more detailed analysis, including examples, is provided in the following subsections to clarify the qualitative appraisal of table 3. Possible temporal feedback between the LCI and LCIA are not assessed, although they may influence LCA results (Weidema et al., 2018).
- 134 Table 2. Meaning of different scores for the qualitative assessment of temporal considerations in LCA

Ranking categories	А	В	С
Relevance	Demonstrated at least in some LCA studies	Expected by authors of this article	Unknown
Conceptual development	A standard method is accepted by the LCA community	At least one method for consideration has been proposed	Theory or concepts have been explained
Operationalisation	Available in the data of most LCA studies when relevant	Some examples have been published	Not found in the literature

136 3.1. Phase of goal and scope definition

In the goal and scope definition, temporal considerations can be introduced by the modelling assumptions, data quality requirements (DQRs) and model limitations. They mostly offer insights on the temporal scope in which LCA studies are representative and useful. This temporal scope also provides an indication of when the dynamic of systems should be considered.

141 3.1.1. Modelling choices

142 Definition of lifetime

143 The lifetime of systems or products, which frames the use phase of the life cycle, is probably the most common 144 temporal consideration in LCA studies (Anand and Amor, 2017; AzariJafari et al., 2016; Fitzpatrick, 2016; Helin 145 et al., 2013; Mehmeti et al., 2016). This temporal scope, which is relative to the overall life cycle, has often been 146 used to ensure a fairer comparison (Joint Research Center, 2010; Jolliet et al., 2010). However, more 147 comprehensive temporal information on the full life cycle, which is not mandatory in international LCA 148 standards (ISO14040, 2006; ISO14044, 2006), would be necessary to explicitly frame the full temporal scope 149 over which elementary flows and impacts might occur. For example, a house can be used for a lifetime of 50 150 years (Hoxha et al., 2016; Standardisation, 2009), but this temporal scope does not include the phase of forest 151 growth, which supplies wood for the fabrication of the building's components (Breton et al., 2018; Fouquet et 152 al., 2015) or for advanced biofuels (Albers et al., 2019a).

153 Dynamic functional units

Some practitioners have suggested that the temporal scope should always be provided with the definition of questions (Finnveden et al., 2009; Huang et al., 2012; Ling-Chin et al., 2016) and functional units (FUs) (Inyim et al., 2016; Santero et al., 2011). The concept of dynamic FUs has been proposed (Kim et al., 2017), which could consider the evolution and comparability of products and would explicitly define the period of validity for a LCA study when the behaviour of consumers and markets have changed significantly. For example, the rapid evolution of technologies for mobile phones has changed their functionalities and demand thus modifying their global production volumes. 161 Table 3: List of temporal considerations in the LCA framework. Rankings for relevance, conceptual development and operationalisation are provided for each consideration on a scale from A to C

162 with their colour code (see table 2). The colour for the three columns of purpose is based on the code of figure 2. The numbers for the rows are the text's subsections.

Sections	Subsection	Temporal considerations	Defining	Considering	Increasing	Relevance	Conceptual	Operationalisation
			temporal	dynamics of	temporal		development	
			scope	systems	representativeness			
	3.1.1 Modelling choices	Definition of lifetime	Х			А	А	А
	5.1.1 Wodening choices	Dynamic FU	Х			А	В	В
		Age of data	Х			А	А	В
3.1 Phase of	3.1.2 Data quality requirements	Technology coverage	Х			А	В	В
goal and scope	(DQRs)	Source of data	Х			А	С	А
definition		Uncertainty description	Х			А	В	В
		Considered life cycle stages	Х			А	А	А
	3.1.3 Limits of assessment	Temporal scope of LCI	Х			А	В	В
		Short- vs Long-term	Х			А	С	В
	3.2.1 Inherent variations	Flows in technosphere		Х		А	В	В
2 2 Dhara af	3.2.2 Temporal resolution	In technosphere		Х		В	В	В
5.2 Phase of	3.2.3 Modelling evolution	Processes in technosphere		Х		А	В	В
System modelling	3.2.4 Prospective modelling	Simulation approaches		Х	Х	В	В	В
System moderning		Historical trends		Х	Х	А	В	В
		Use of scenarios		Х	Х	А	В	В
3.3 Phase of inventory analysis:	3.3.1 Framework	Matrix-based		Х		А	В	В
		Graph traversal		Х		А	В	В
	3.3.2 Approach and tool	DyPLCA		Х		А	В	В
		Temporalis		Х		А	В	В
	3.4.1Modelling choices	Time Horizons	Х			А	А	А
		Discounting	Х			С	В	С
		Period of validity	Х			В	В	В
	3.4.2 Limits of assessment	Short- vs Long-term	Х			А	С	В
3.4 Phase of impact assessment	3.4.3 Temporal indicator	Payback time	Х			В	В	В
	3.4.4 Inherent variations	Non-linear mechanisms		Х	Х	В	В	С
	3.4.5 Temporal resolution	Ecosphere mechanisms		Х		В	С	С
	3.4.6 Modelling evolution	Background concentration		Х	Х	В	В	С
	3.4.7 Prospective modelling	Scenarios		Х	Х	В	В	В
		Period-specific CFs		X	X	В	В	В
	3.4.8 Computational framework	Characterisation functions		X	X	C	C	C
		DvPLCA		Х		А	В	В
	3.4.9 Approach and tool	Temporalis		Х		А	В	В

164 3.1.2. Data quality requirements (DQR)

165 Age of data

Some metadata of datasets, which should be defined in the DQR (ISO14044, 2006; Joint Research Center, 2010), informs on their age and minimum length of time for data collection. Potential temporal discrepancies between used datasets and the targeted temporal scope of a modelled system can thus be partially evaluated. Such information also provides some insights on the temporal scope of a system model when it represents human activities (Bessou et al., 2013; Yuan et al., 2015). For example, the description of solar energy installations from the 1990s would probably be relevant for LCA of solar energy before 2000. Nevertheless, using such periods of validity require expert opinion, thus limiting the usefulness for this kind of metadata.

173 Technology coverage

174 In some cases, the definition of technology coverage in the DQR of datasets can inform on the actual temporal scope of the study (ISO14040, 2006; ISO14044, 2006; Joint Research Center, 2010) with the ensuing qualitative 175 176 assessment of temporal representativeness. For example, ecoinvent (Wernet et al., 2016) uses five levels of 177 technology (i.e. new, modern, current, old and outdated) to describe transforming activities. Using datasets with 178 new or modern technology levels should therefore be relevant for LCA studies on future products. However, this 179 information is relative to each sector, as the modern level could be representative for 10 years of technology 180 evolution in an established sector, whereas fast-paced sectors like electronics may use modern technologies for 181 only 1 year before switching to new options.

182 Source of data

183 The choice of data sources and the qualitative assessment of their overall representativeness provide an indirect assessment of the temporal scope for modelled systems and LCA studies (Rebitzer et al., 2004). For example, 184 185 when data are sourced from scientific journals, date of publication is the primary indication for its period of 186 validity. More precise temporal information is also often provided in case studies for systems with longer lifetimes or in DLCA studies like (Heeren et al., 2013; Pahri et al., 2015; Sohn et al., 2017a; Vuarnoz et al., 187 2018). The use of up-to-date LCA databases can bring a false sense of security on the temporal scope and 188 representativeness of the data for recent products or systems. Indeed, database updates do not always follow the 189 190 changes in market shares or evolution of technology because of the lack of new data.

191 Nevertheless, different temporal metadata is given for most datasets. For instance, ecoinvent guidelines (Wernet 192 et al., 2016) require the definition of the date of generation, the date of review and the period of validity with a 193 start date and end date for any dataset. These temporal considerations fulfil most of the requirements from ISO 194 14044 (2006) except for the definition of the averaging period of dataset inputs. The ILCD handbook (2010) has 195 set further requirements defining temporal properties: the expiring year of datasets and the duration of the life 196 cycle, which respectively relates to the period of validity for LCI datasets and the temporal scope of elementary 197 flows for a dataset. This metadata is available in most datasets of the ELCD (Recchioni et al., 2013). Many of 198 these temporal metadata are more relevant to assess the temporal scopes of studies than the choice of a database 199 and its version, but the place (e.g. in dataset descriptions) and the different definition under which they can be 200 found hinder their use in most LCA studies.

201 Uncertainty description

202 The description of the uncertainty associated with flows (e.g. in econvent (Wernet et al., 2016)) is another 203 indirect source of information to clarify the temporal scope and period of validity. Indeed, the temporal 204 correlation indicator provides a quantitative assessment of the discrepancy between the time when the data was 205 acquired and the intended temporal scope for the dataset (Weidema et al., 2012). For example, a product flow 206 with a temporal correlation indicator of 3 means that its value has been gathered between 6 and 9 years before or 207 after the targeted temporal scope of the dataset. With the current definition of the temporal correlation indicator, 208 the precision of this temporal information is rather low (i.e. >3-year period) and is widely missing in LCA 209 databases and studies, limiting its applicability.

210 3.1.3. Chosen limits of assessment

The definition of limitations in the stage of goal and scope definition is probably the step where temporal scopes are defined with higher precision and clarity in LCA studies, even more in recent DLCA studies. While this is useful, typical LCA reports mainly offer qualitative definitions, which are not sufficiently transparent to describe the considered period in assessed life cycles.

215 *Considered stages of the life cycle*

LCA studies can limit the temporal scope of their analysed systems and LCIs by considering only a part of the life cycle. Setting the end-of-life outside the boundaries is an example of such a limited temporal scope. The ISO 14044 (2006) allows this limitation, but only if they do not significantly change the overall conclusions of a study because such phases are not linked to significant impacts. Most of the LCA reports clearly state the excluded life cycle stages, but they often provide an imprecise description for the limitation of the temporal scope. Moreover, the specification of the considered stages of a life cycle will not explicitly state the temporal scope in which elementary flows are considered (e.g. 2 years) nor offer a calendar-based period of occurrence (e.g. from January 2019 to December 2020).

224 Temporal scope of life cycle inventories

225 More specific and precise descriptions of temporal scopes for LCI have been provided in recent scientific publications that focus on some temporal considerations (i.e. DLCA). For example, relative temporal scopes 226 227 have been used to define the periods of LCIs for many studies on different products, for example considering the lifetime of wood-based products and buildings between 50 and 100 years (Fouquet et al., 2015; Levasseur et al., 228 229 2010) including tree growth period over 70 and 150 years (Levasseur et al., 2013; Pinsonnault et al., 2014), lifetime of marine photovoltaic of 20-30 years (Ling-Chin et al., 2016) and zinc fertiliser over 20 years crop 230 231 rotation (Lebailly et al., 2014). In these cases, the LCIs are enclosed within a quantified period of time that can 232 be relevant for some impact categories, but they lack any reference to a calendar year or period. Several DLCAs 233 studies defined calendar-based temporal scopes, but discussions on the potential usefulness of this contextual 234 information could be further enriched. Some were based on reference calendar years of building materials 235 (Collinge et al., 2013b), hourly energy demand in buildings (Vuarnoz et al., 2018), as well as seasonal and 236 annual variations in crop rotations (Caffrey and Veal, 2013). Other studies were based on calendar-specific 237 periods detailing domestic hot water production (Beloin-Saint-Pierre et al., 2017), future biomass production (Menten et al., 2015), the lifetime of buildings (Roux et al., 2016a; Roux et al., 2016b), the energy use in hourly, 238 239 daily and monthly temporal resolutions (Collinge et al., 2018; Karl et al., 2019), or for introducing back-time 240 horizon (Tiruta-Barna et al., 2016).

241 Short- vs long-term analysis

Several publications describe the temporal scopes of technosphere models (Dandres et al., 2012; Menten et al., 2015) or LCI (Finnveden et al., 2009; Morais and Delerue-Matos, 2010; Pettersen and Hertwich, 2008; Roder and Thornley, 2016) with adjectives such as short-, medium- or long-term. These qualitative and relative attributes thus inform the considered periods, but are vague. This lack of a precise temporal definition can be partly explained by the lack of consensus on how temporal scopes should be defined.

248 3.2. Phase of inventory analysis: system modelling

In the system-modelling step of the LCI phase, temporal considerations are found in the descriptions of the system inherent variations and evolution. They define the dynamics of systems and can improve the temporal representativeness of models for technosphere activities (i.e. network of processes). Although considering system evolution and inherent variations in both the foreground and the background data is still not a common practice, its importance has long been acknowledged in ISO 14040 (2006), stating that "*all significant system variations in time should be considered to get representative results*".

Strategies to consider inherent variations and evolution have been proposed by different authors, mainly for energy (Amor et al., 2014; Zaimes et al., 2015), transport (Tessum et al., 2012), agriculture (Fernandez-Mena et al., 2016; Yang and Suh, 2015) and waste management (Bakas et al., 2015). For example, the energy share of electricity production in a country varies throughout days, weeks, months and seasons (Beloin-Saint-Pierre et al., 2019; Vuarnoz and Jusselme, 2018). LCA case studies have shown that inherent temporal variations of production can have significant effects on results, mainly when consumption of these products is not constant over time.

262 3.2.1. Inherent variations with flow differentiation

263 Inherent variations can be modelled with temporal differentiation of flows or dynamic modelling. For instance, electricity production (Messagie et al., 2014; Vuarnoz and Jusselme, 2018; Walker et al., 2015) and its use in 264 buildings (Collinge et al., 2013b; Collinge et al., 2018; Karl et al., 2019; Roux et al., 2016b; Roux et al., 2017; 265 Vuarnoz et al., 2018; Walzberg et al., 2019a), cloud computing (Maurice et al. 2014) and wastewater treatment 266 267 (de Faria et al. 2015) have all been modelled with such approaches. In different ways, all these approaches 268 convert flows into temporal distributions, thus supplementing temporal properties to the core data of the model 269 components in the LCA framework. The applicability of such data in other LCA studies is often limited because 270 the temporal information is valid only for the temporal scope of a given case study. A way to address this limitation is to use a reference "time 0" in the temporal distribution as a period of occurrence relating to a 271 starting period of a process (Beloin-Saint-Pierre et al., 2014; Tiruta-Barna et al., 2016). This "time mark" creates 272 273 process-relative descriptions, which can be reused in any period of a life cycle or even for different life cycles. 274 Tiruta-Barna et al. (2016) and Pigné et al. (2019) provided process-relative temporal distribution archetypes for 275 ecoinvent v3.2, applicable to foreground and background datasets. As underlined by Beloin-Saint-Pierre et al. 276 (2014), the additional efforts needed to provide temporal information for all the flows of LCA databases are still significant and the prioritisation of data-gathering remains important. 277

278 3.2.2. Temporal resolution

279 The level of temporal resolution to models the dynamics of systems depends on the sector and the modelling 280 approach. For instance, hourly resolutions have been chosen for electricity production and consumption (Amor et al., 2014) or the transportation sector (Tessum et al., 2012). For assessing long-term emissions, for instance 281 282 from waste treatment, a temporal resolution of centuries is more appropriate (Bakas et al., 2015). Some authors 283 have proposed a temporal differentiation based on archetypes. For example, archetypal weather days (Risch et 284 al., 2018) have been developed to contrast the relative importance of episodic wet weather versus continuous 285 dry-weather loads. So far, studies about the consequences for choosing different temporal resolutions to describe 286 the flows are limited. Indeed, only two examples are found in the building sector where a monthly resolution is 287 deemed sufficient to consider most of the temporal variability (Beloin-Saint-Pierre et al., 2019; Karl et al., 2019).

288 3.2.3. Modelling evolutions with process differentiation

The basic strategy to describe evolution is to differentiate processes when a system is considered to change substantially over time. The key challenge here is to identifying when changes are significant enough without expert opinion on the modelled product. A simple application can be performed, if calendar-based periods of validity are consistently provided for all datasets in LCA databases; they could then be changed automatically when they are no longer valid representations over the full life cycle of any system. Such metadata is, however, required only in the (discontinued) ELCD database (see subsection 0) and, currently cannot be easily integrated in LCA software.

Collet et al. (2011) proposed an approach to tackle this problem and identify where temporal differentiation of processes during system modelling is needed. Their general idea is to recognise when the combined emission and impact dynamics justify the additional effort for temporal differentiation. Moreover, the selective introduction of the time dimension in background processes has been studied by Pinsonnault et al. (2014) and more recently by Pigné et al. (2019). The authors have shown that the temporal variations of a selection of background processes and the entire ecoinvent database can significantly affect climate change impacts for processes in some sectors (e.g. transport and building).

303 3.2.4. Prospective modelling

304 Modelling future evolution of systems is another common example of temporal considerations that is often 305 performed under the umbrella of DLCA studies. Indeed, many DLCA studies have explored different 306 prospective models for a range of products like: photovoltaic panels (Pehnt, 2006; Zhai and Williams, 2010), 307 buildings (Collinge et al., 2013a; Frijia et al., 2012; Scheuer et al., 2003; Sohn et al., 2017a; Sohn et al., 2017b; 308 Su et al., 2017), bioethanol (Pawelzik et al., 2013), passenger vehicles (Bauer et al., 2015; Miotti et al., 2017; 309 Simons and Bauer, 2015), metals (Stasinopoulos et al., 2012) or ammonia (Mendivil et al., 2006). Any temporal 310 assumptions made to define future evolution are thus considered for system modelling and LCI calculations. 311 While major advances have been reached to offer explicit descriptions of assumptions made for temporal 312 considerations in DLCA, e.g. (Collinge et al., 2013b; Herfray and Peuportier, 2012; Menten et al., 2015; Pehnt, 313 2006; Roux et al., 2016b), they are currently not the standard. Prospective modelling assumptions can be 314 grouped within three categories that have fundamental differences on how they justify their forecasting.

315 Simulation approaches

316 Economic models, such as partial equilibrium models (PEM) or general equilibrium models (GEM), are 317 frequently used in, but not limited to, consequential LCA modelling to simulate potential future evolution to 318 assess direct and indirect consequences of decisions (e.g. climate policies) on large scale systems. Nevertheless, 319 the current focus of using these models to assess consequences of changes in LCA studies should not hide their 320 potential to offer possible development paths in prospective assessments. PEM generally focuses on one 321 particular economic sector with a higher level of detail (i.e. technology rich), while GEM covers the whole 322 economy with a lower level of detail (typically 30-50 economic sectors). For instance, PEMs have been used to model the energy sector in France (Albers et al., 2019c; Menten et al., 2015), or biogas production in 323 324 Luxembourg (Marvuglia et al., 2013) and GEMs have been used to evaluate the consequences of different 325 energy scenarios on the whole economy in Europe (Dandres et al., 2011). PEMs have also been coupled with 326 GEMs to model the consequences of energy policy scenarios in an integrated manner (Igos et al., 2015) and they 327 have been used in combination with dynamic models of biogenic and soil organic carbon for a similar purpose 328 (Albers et al., 2020; Albers et al., 2019b).

The lack of consideration for human behaviour in PEM or GEM has recently been pointed out as a potential issue for the validity of the prospective models (Marvuglia et al., 2015). The use of agent- or activity-based models have therefore been proposed as alternatives to carry out prospective assessments; both in the foreground and in the background systems. Such models have mostly been used in consequential LCAs relating with transport policies (Querini and Benetto, 2015), regional market penetration of electric vehicles (Noori and Tatari,
2016), switch grass-based bioenergy systems (Miller et al., 2013), smart buildings (Walzberg et al., 2019b) or
raw materials criticality (Knoeri et al., 2013), but could be used to predict future trends. The differences between
the use of such simulation approaches in DLCA or consequential LCA studies have been discussed recently by
Sohn *et al.* (2020).

338 Forecasting based on historic trends

339 Some data sources (e.g. statistics on energy production) describe historic trends from which forecasting is made 340 by extrapolation, assuming paradigm shifts will not occur. For instance, regression analysis was used to assess 341 the evolution of energy systems (Pehnt, 2003a; Pehnt, 2003b; Pehnt, 2006; Yang and Chen, 2014) and the 342 construction sector (Sandberg and Brattebø, 2012). The main strength of this approach is its simplicity and the 343 potential to assess the observed level of variability of historic trends. It can thus provide averaged future trends 344 and the expected variability (uncertainty). The main weakness, on the other hand, is the implicit assumption that 345 historic trends are representative of the future, which is not always the case, particularly for emerging systems 346 and technologies.

347 Using scenarios to explore potential futures

348 Scenario-based modelling has been used in many sectors like waste management (Hellweg et al., 2005), water 349 consumption (Pfister et al., 2011), bioenergy (Choi et al., 2012; Daly et al., 2015; Dandres et al., 2012; Earles et 350 al., 2013; Igos et al., 2014; Menten et al., 2015), renewable electricity (Hertwich et al., 2015; Pehnt, 2006; 351 Viebahn et al., 2011), transport (Cheah and Ieee, 2009; Garcia et al., 2015; Pehnt, 2003a; Pehnt, 2003b), 352 chemicals (Alvarez-Gaitan et al., 2014) and buildings (Roux et al. 2016b). A general idea behind modelling 353 scenarios is that exploring many potential futures may be simpler to justify than offering predictions on what the 354 future will look like for a system as complex as human activities. For instance, Pesonen et al. (2000) defined that 355 the scenarios describe possible future situations based on assumptions about the future and include developments from the present to the future. The authors distinguished between "what-if" and "cornerstone" scenarios 356 357 (Pesonen et al., 2000), depending on the need to consider short- or long-term planning. "What-if" scenarios are 358 often based on the field-specific expertise of LCA practitioners. Cornerstone scenarios explore many options 359 with very different assumptions on the future to identify potential development paths. Another category is legally 360 bound scenarios that explore future paths under the restriction of regulations.

361 3.3. Phase of inventory analysis: LCI computation

The computation of LCI transforms the information of a technosphere model into a set of elementary flows whose quantities are in relation to the FU of the assessed systems. The computation traditionally aggregates all flows of the same type over the entire life cycle.

365 3.3.1. Computational framework

366 Matrix-based computation with process differentiation

367 The conventional matrix-based computational approach can be used to calculate DLCIs, but with larger technosphere and ecosphere matrixes (Heijungs and Suh, 2002). Collinge et al. (2012; 2013b) used this approach 368 369 on foreground processes to calculate the DLCI for each year of a building's life cycle. They concluded, similarly 370 to Heijungs and Suh (2002), that the implementation brings significant challenges in data management when 371 background databases are used. The challenges of this approach are twofold. Firstly, the temporal description of 372 a system needs to be re-informed when the periods of assessment differ (e.g. 1980-2000 vs 2005-2025), if 373 considered impacts are calendar-based. Secondly, the amount of data and the computational efforts depend on 374 the required temporal precision (e.g. day vs. year) to describing all flows.

375 Graph traversal structure

The Enhanced Structure Path Analysis (ESPA) approach (Beloin-Saint-Pierre et al., 2014) is one type of graph-376 377 based computational framework that convolves process-relative temporal distributions (see subsection 3.2.1) to 378 propagate the temporal descriptions of flows. The general concept behind the ESPA framework (Beloin-Saint-379 Pierre et al., 2014; Maier et al., 2017) relates to one strategy of graph traversal algorithms (i.e. breadth-first), but 380 other options have been explored. The depth-first search strategy (Tiruta-Barna et al., 2016) recommends a 381 different traversal of supply chains, which is normally linked to lower memory requirements. The best-first 382 search strategy (Cardellini et al., 2018) is another option that propagates the temporal information by prioritising 383 the temporal distribution with higher contributions to impacts. All these options use process-relative temporal 384 distributions, thus profiting from their reusability and the potential for higher temporal precision.

385 3.3.2. Approaches and tools

Some commercial software tools use matrix-based computation (e.g. Simapro, Umberto) and could thus work with the process differentiation framework for the calculation of temporally differentiated LCI. To our knowledge, this option has not been implemented comprehensively in DLCA studies because LCA databases do not offer temporal details. The ESPA method has also not been developed into a computational tool and its implementation has been limited to one simplified case study (Beloin-Saint-Pierre et al., 2017). Nevertheless,

391 two options currently exist for full DLCI computations and are introduced in the following sub-sections.

392 DyPLCA

393 DyPLCA has been implemented as a web tool (available at http://dyplca.univ-lehavre.fr/), originally presented 394 by Tiruta-Barna et al. (2016), which uses the depth-first graph search strategy. The main parameters that 395 balance accuracy vs. computation time in this tool are the temporal resolution of function integrals and 396 the back time span. Common values for both are respectively 1 day and -50 years (i.e. 50 years before the 397 period of occurrence for the FU). The computational intensity of the DLCI calculation has thus been 398 resolved by a trade-off between accuracy and cut-offs. The process-relative temporal distributions can 399 have different levels of detail to describe the flows in the system models. For instance, they can be detailed 400 for foreground processes, as presented in Shimako et al. (2018), and can be rather generic for the 401 background datasets.

402 DyPLCA currently works with a temporal differentiated ecoinvent v3.2 (Pigné et al., 2019), providing generic
403 temporal descriptions to most background inventory processes. The DLCI results can be further used with static
404 or DLCIA methods, as shown in studies on bioenergy production from microalgae (Shimako et al., 2016) and on
405 grape production (Shimako et al., 2017).

406 Temporalis

407 Temporalis (Cardellini et al., 2018) is a free and open source package of the Brightway2 LCA tool (Mutel, 408 2017), using the best-first search strategy. The tool is fully compatible with many existing commercial LCA 409 databases, but temporal descriptions of datasets are currently not provided. Temporalis does not require a fixed 410 and continuous temporal resolution over any system models to provide DLCI or results for the impact 411 assessment. Nevertheless, a DLCIA method for GWP based on the IPCC methodology (2013), is included. A 412 simple case study for the temporal consideration of biogenic carbon flows was carried out with the method of 413 Cherubini et al. (2012; 2011). It has shown that the LCI computation can be resolved on a regular laptop within a 414 short time. Nevertheless, further developments still need to be completed before most LCA practitioners can use 415 the tool easily.

416 3.4. Phase of life cycle impact assessment

In the LCIA phase, temporal considerations affect many aspects that are linked to all phases of the LCA
framework. For instance, the selection of a TH and changes of environmental mechanisms (i.e. impact pathways)
over time are key modelling choices to characterise impacts in a DLCA framework.

420 3.4.1. Modelling choices

LCIA is a complex task that requires many assumptions (e.g. the future state of the environment) and choices, which sometimes limit the validity of results to a specific temporal scope and introduce bias in the results. One of the most explicit and commonly used temporal considerations in LCIA methods is the TH, restricting the impact assessment to a specific period. Discounting is another modelling choices that can affect LCA results in similar ways to TH with links to its potential subjectivity (Lueddeckens et al., 2020).

426 Time Horizon

427 The choice between a finite or infinite TH is a common type of temporal consideration that sums the 428 environmental effects over a selected temporal scope (e.g. the 100-year TH for the GWP indicator). The 429 consideration of different THs is used, for instance, by the ReCiPe method (Huijbregts et al., 2016), which builds on three cultural perspectives, proposed by Hofstetter et al. (2000). These perspectives are associated with 430 431 different sets of calculation assumptions, including CFs with different THs for each impact category. For 432 example, the "hierachist" perspective retains a 100-year TH for GWP and other categories, while "individualist" 433 and "egalitarian" perspectives respectively use THs of 20 and 1000 years. Furthermore, very long THs are 434 suggested for some impact categories such as for climate change (i.e. 1000 years) and ionising radiation (i.e. 435 100,000 years). The ILCD handbook (2011) and the SimaPro Database Manual (PRé, 2016) provide additional 436 insights into the use of THs in different LCIA methods, but there is not yet any standard on how to deal with 437 long-term impacts and related uncertainties within all categories. For instance, the 5th IPCC assessment report 438 (2014) removed the 500-year TH due to high uncertainties associated with the assumption of constant 439 background concentrations.

To date, the choice of a TH remains a topic of discussion within the LCA community (Dyckhoff and Kasah,
2014; Reap et al., 2008b) where three critical aspects are challenging the use of fixed and finite THs in LCIA
methods:

• The first aspect is the inconsistency between the temporal boundaries of the studied systems and the TH of the LCIA methods (Benoist, 2009; Levasseur et al., 2010; Rosenbaum et al., 2015; Yang and Chen, 2014).

445 Indeed, it could be understood that effects from elementary flows beyond the chosen TH should not be 446 considered. However, the effects are ultimately modelled over an invariable temporal scope, even if they occur at different periods during a life cycle (e.g. 100 years). This use of THs may thus lead to 447 448 misrepresentations of impacts and their period of occurrence (Hellweg and Frischknecht, 2004), for 449 instance, misleading decision-making concerning temporary storage and emission delays (Brandao and 450 Levasseur, 2011; Jørgensen et al., 2015). This issue can be particularly significant for intermitting emissions like pesticides, where arbitrary cut-offs of emissions after pesticide application should influence how each 451 452 emission contributes to related impacts of human toxicity (Fantke and Jolliet, 2016) and ecotoxicity (Peña et al., 2019). 453

454 The second aspect refers to the time integration of substances with highly variable environmental effects 455 over their lifetime in the ecosphere (e.g. aging effects reducing bioavailability of metals (Owsianiak et al., 456 2015) or transformation of persistent chemicals in the environment (Holmquist et al., 2020)), which can 457 significantly bias the conclusions of LCA studies (Arodudu et al., 2017; Lebailly et al., 2014). In the case of 458 GWP, the weight of forcers with very short atmospheric residence time decreases with an increasing TH (Levasseur et al., 2016; O'Hare et al., 2009), while a shorter TH increases the importance of short-lived 459 gases. For example, methane (CH_4), whose atmospheric lifetime is about 12.4 years, goes from a factor of 460 84 CO₂-eq for the 20-year TH to a factor of 28 CO₂-eq for 100-year TH (Myhre et al., 2013). For further 461 462 examples on this subject, Levasseur et al. (2016) presented various approaches that have been proposed for TH definition. For toxic substances, Huijbregts et al. (2001) demonstrated that TH variations can change 463 impacts by up to 6.5 orders of magnitude for metal toxicity. In this case, the high dependency between CFs 464 465 and the chosen TH is due to long residence times (i.e. persistence) in fate models, which increase metal run-466 offs and leaching potential to global marine and soil compartments.

The third aspect relates to the temporal cut-offs that come with the selection of a fixed and finite THs, which 467 can be ethically questioned in the context of intergenerational equity (Hellweg et al., 2003a). Indeed, these 468 cut-offs raise concerns on the subjectivity of choosing a specific TH to highlight preferences between short-469 470 and long-term impact considerations (Lueddeckens et al., 2020). For instance, the 100-year TH in GWP is the most used and recommended option, but this preference is not justified by scientific facts (Reap et al., 471 2008b; Shine, 2009; Vogtländer et al., 2014) and is implicitly subjective for decision-making (Brandao and 472 473 Levasseur, 2011; Fearnside, 2002). This 100-year TH is particularly important when temporary/permanent 474 carbon storage or the delayed emissions from biogenic and fossil sources are evaluated or incentivised

- 475 (Guest and Stromman, 2014; Levasseur et al., 2012a). Moreover, emissions that are delayed after the 100-
- 476 year scope are then considered to be permanently avoided (BSI, 2011; Joint Research Center, 2011).

478 A "simple" solution to remove such time preferences and value choices has been recommended by setting infinite THs in all cases. For instance, some LCIA methods (e.g. EDIP2003 (Hauschild et al., 2006), IMPACT 479 480 2002+ (Jolliet et al., 2003), ReCiPe 2016 (Huijbregts et al., 2016)) use infinite or indefinite THs as a standard for 481 stratospheric ozone depletion, human toxicity and ecotoxicity. In the case of the land use impact category, THs 482 are generally not explicitly stated in current characterisation models (see e.g. Huijbregts et al. (2016) for 483 biodiversity impacts or Müller-Wenk and Brandão (2010) for climate change). Even if the theoretical 484 frameworks for land use impact assessment discusses changed (Beames et al., 2015) or permanent impacts and 485 therefore the need for defining a TH (Canals et al., 2007; Koellner et al., 2013), permanent impacts are currently 486 not considered in available characterisation models. Current models implicitly correspond to the choice of an 487 infinite TH where impacts of each land use intervention is being integrated over time until the effect factor 488 reaches 0, i.e. until the variations of soil quality after the land use intervention regenerates back to a reference 489 soil quality. Regeneration time then plays a significant role in the effective integration period and in the 490 definition of CFs.

491 Discounting

This concept was discussed to value time in LCIA (Hellweg et al., 2003a; Pigné et al., 2019; Yuan and Dornfeld,
2009; Zhai et al., 2011) and to deal with the uncertainties associated with time preferences and future emissions.
The setting of finite THs is an implicit form of discounting for long-term impacts, using a zero discount rate over
the TH, and an infinite discount rate beyond the TH. Discounting offers a trade-off between giving a higher
value to present or future impacts. A more detailed discussion on this subject is provided by Lueddeckens et al.
(2020).

498 3.4.2. Chosen limits of assessment

The periods of validity for chosen LCIA methods and discussions on the short- or long-term nature of impacts are two types of temporal considerations that can inform on the temporal scope of a LCA study, whether this selection is voluntarily made by the practitioner or not.

502 Period of validity for LCIA methods

503 Stating the period of validity (e.g. 2000 to 2010) or version for chosen LCIA methods in LCA studies is not 504 common practice, but it can provide insights on the expected temporal scope (Bessou et al., 2011; Hauschild et 505 al., 2013; Ling-Chin et al., 2016; Weidema et al., 2012). The choice of THs can also suggest an implicit 506 definition of the considered period of validity. In an ideal world, the temporal scope of obtained LCIs and chosen 507 LCIA methods should be fitted to each other. Such a correspondence is desirable if CFs vary significantly over 508 time, but it is currently difficult to implement in the available databases and software tools.

509 Short- vs long-term analysis

510 Much like it has been said in the definition of the goal & scope (subsection 3.1.3), the adjectives of short- and 511 long-term have been used to describe the temporal scope of LCIA methods (Arodudu et al., 2017; Chowdhury et 512 al., 2017; Reap et al., 2008b). This lack of a precise temporal definition when stating short-, medium- and long-513 term can be partly explained by the differences in time scales of life cycles and environmental impacts for 514 different systems. Furthermore, a commonly accepted standard does not yet exist to deal with long-term impacts 515 and related uncertainties within all categories. For instance, the 5th IPCC assessment report (Myhre et al., 2013) 516 removed the previously published 500-year TH due to the high uncertainties associated with the assumption of 517 constant background concentrations.

518 3.4.3. Temporal indicator

519 Payback time

Payback times have been created to provide a temporal scope that informs on temporality of impacts. The basic idea is to calculate the necessary period to compensate for the "cradle-to-gate" impacts of any system. It has been mostly used to evaluate the time it takes to produce an amount of electricity that is equivalent to the primary energy use from the manufacturing of photovoltaic installations (Espinosa et al., 2012; Fthenakis and Alsema, 2006; Knapp and Jester, 2001), but it can be applied to energy use in many types of product (Elshout et al., 2015) or could also give payback time for other impact categories.

526 3.4.4. Inherent variations

In conventional LCIA methods, CFs are determined with average or marginal approaches that model changes in the impact according to a change in the inventory (Frischknecht and Jolliet, 2016; Hauschild and Huijbregts, 2015). With this average approach, the environmental disturbances from different activities are aggregated, historically referred to as "snapshots" of a studied system (Bright et al., 2011; Heijungs and Suh, 2002; Klöpffer, 2014; Levasseur et al., 2016; Owens, 1997b; Vigon et al., 1993). For example, most existing models for characterising toxic impacts (Rosenbaum et al., 2008) assume constant environmental conditions for the assessment of health impacts. With this approach, inherent variations of the ecosphere are not considered.

534 Non-linear mechanisms in the ecosphere

The marginal approach addresses an impact resulting from a small change to a given background concentration. The impact is therefore positioned in relation to the current environmental state. For example, studies of human health impacts from exposure to fine particulate matter ($PM_{2.5}$), where indoor, outdoor, urban and rural locations have shown significant differences in $PM_{2.5}$ background levels (Fantke et al., 2017). A non-linear exposureresponse model thus accounts for these differences in $PM_{2.5}$ levels, reflecting a slope for low concentrations that are substantially higher than for high concentrations (Fantke et al., 2019).

541 Impact assessment models are representations of complex environmental mechanisms that depend on a long list 542 of parameters, such as the lifetime of substances in the environment and the sensitivities of ecosystems over 543 different temporal scopes (Lenzen et al., 2004). In many LCIA methods, CFs are defined from generic 544 parameters values in stationary conditions (e.g. intervention quantity, baseline for target substances, and profiles 545 of the soil composition) or for a given TH. Subsequently, impacts are assumed linearly proportional to the 546 inventoried emissions, which enable the scaling of impacts to any functional unit. In reality, the involved 547 environmental mechanisms are dynamic and often highly complex (Arbault et al., 2014). They depend on the 548 physical, chemical and biological phenomena and non-linear interaction occurring in nature and are 549 consequences of the elementary flows generated by human activities.

550 Time-dependent characterisation has been performed in some cases by modelling the dynamics for one or more 551 of the three factors influencing an impact (i.e. environmental fate, exposure, and effects), thus creating a type of 552 DLCIA methods. Effect data are typically not easily linked to temporal properties, allowing for temporal 553 considerations in effect modelling (e.g. dose response for human effects or concentration response for ecological 554 effects). Hence, time-dependent characterisation is usually only facilitated by considering the dynamics of 555 systems in the fate and exposure factors of an impact pathway, which is usually enabled by models of the 556 underlying mass balance for a given impact pathway. This has been implemented, for example, in toxicityrelated impacts (Lebailly et al., 2014), where the system dynamics of the environmental fate factor are either 557 solved via numerical integration (Shimako et al., 2017), or via matrix decomposition (Fantke et al., 2013). 558

559 3.4.5. Temporal resolution

560 Specific temporal resolution for each elementary flow

561 The temporal considerations within LCIA models may follow specific frequencies (*e.g. yearly changes*), as well 562 as temporal-inherent features deriving from dynamic biogeochemical processes. The frequency can be

differentiated, for instance, as responding to episodic (e.g. initial land clearing), cyclical (e.g. seasonal water and 563 564 pesticide use), stochastic (e.g. 1 in 20 years' waste discharge), or continual (e.g. fisheries yields) variations in the 565 studied system (Lenzen et al., 2004). Cyclical or seasonal variations concerning sunlight, temperature and 566 precipitation on the calendar year (e.g. winter vs summer time) are other examples of temporal considerations 567 that could be relevant for impact categories like aquatic eutrophication (Udo de Haes et al., 2002), water scarcity 568 (Boulay et al., 2015), human toxicity (Manneh et al., 2012) and photochemical oxidant formation (Shah and 569 Ries, 2009). Such frequencies therefore highlight relevant temporal resolutions for the temporal differentiation of 570 elementary flows in databases and DLCIs. Temporal inherent features may vary with hourly, daily, monthly or 571 yearly constraints depending on temporal patterns or modelling time steps of the characterisation models (Collet, 572 2012; Owens, 1997b).

573 The temporal scope of impact assessment itself may be aligned with the dynamics of governing biogeochemical 574 processes to more accurately represent certain fate dynamics. For instance, Liao et al. (2015 found that common 575 seeding-to-harvest assessment periods in agricultural LCAs do not correspond to the actual dynamics of 576 fertilising substances, some of which contribute to eutrophication during the next crop rotation. The same 577 concerns agricultural pesticides, where the time between the application and crop harvest drives related residues leading to human exposure (Fantke et al., 2011). Such fate dynamics can still be analysed and parameterised to 578 579 fit steady-state models and associated impact pathways, such as human toxicity (Fantke et al., 2012; Fantke et 580 al., 2013).

581 3.4.6. Modelling evolutions

582 Considering variations for concentration substances and the state of the environment

583 Elementary flows may have varying levels of effect, depending on the timing of emissions (i.e. period of 584 occurrence) and the state of the environment (i.e. varying substance concentrations). Temporal considerations of 585 environmental mechanisms in LCA studies are challenging because the current state of practice rarely allows to 586 account for the periods of emission occurrences that are related to a product's life cycle (Finkbeiner et al., 2014; 587 Hellweg and Frischknecht, 2004; Jørgensen et al., 2014; Kendall et al., 2009; Levasseur et al., 2010; Reap et al., 588 2008b). In fact, LCI flows are typically given as simple values that are considered to be a representation of 589 steady or pulsed flows from and to the environment by most LCIA models. For instance, impacts 590 characterisation methods often use an effect factor for a given concentration of pollutants in the background 591 environment (Finnveden et al., 2009; Hauschild, 2005). Thus, the same amount and type of elementary flows 592 (*i.e.* equivalent LCIs) can generate different levels of impacts because they have been emitted at different periods

593 of occurrence (e.g. 2016 or 2017), with varying flows (i.e. inherent variations) and geographies, requiring both 594 temporal and spatial differentiation. In this case, calendar specifications may be relevant to assess and compare 595 the evolution of impacts and/or background concentrations over time (e.g. 1990 Kyoto Protocol and the 1750 596 IPCC reference years for climate change). The inherent variations in the state of the environment can also affect 597 the CFs. For example, temporary changes in the carbon cycle from land use (Vazquez-Rowe et al., 2014) and 598 related changes in the albedo of the land surface are two dynamic aspects that can bring variations in 599 environmental impacts (Bright et al., 2012). Such variations are currently difficult to assess since they are not 600 linked to "standard" elementary flows, which are always the source of impacts in the usual LCA framework.

601 3.4.7. Strategies for prospective modelling

As is the case for technosphere models, it is, in principle, possible to forecast the environmental responses of the ecosphere to elementary emissions with the use of scenarios.

604 Scenarios

An alternative form of temporal considerations in LCIA is increasingly performed on scenario-driven case studies. It has been applied to water use impacts by means of scenario-bound CFs, where each scenario represents a different prospective TH (Núñez et al., 2015). It is a step towards considering the temporal variability of environmental indicators, as most LCIA methods make the implicit assumption that the environment and its properties will not evolve over the studied life cycle. Another common example is the case of metal leaching in ground that has been forecasted with different scenarios (Huijbregts et al., 2001; Pettersen and Hertwich, 2008).

612 3.4.8. Computational framework

Recently, some DLCIA methods have been developed with different computational frameworks. These approaches are key to understand the links between DLCIs and DLCIA methods, while offering potential pathways for future developments.

616 Period-specific characterisation factors

In the last decade, LCA researchers have developed DLCIA methods addressing time dependent impacts as a function of time, yet they are mainly restricted to GWP and toxicity indicators. These DLCIA methods consider the periods of occurrence for emissions by providing different period-specific CFs to assess their impacts. For example, CFs can be calculated for each year over a chosen time horizon or for the month of January 2020. These CFs thus bring consistency between the temporal scopes of DLCI and impacts (Levasseur et al., 2010). 622 Different LCA scholars found that the results based on such DLCIA methods provide useful examples for 623 decision-making, among others, on: "the intensity, extend and frequency of the impacts" (Lebailly et al., 2014), 624 the sensitivity of the results to various TH choices (Levasseur et al. 2012b), and the optimisation options from 625 scenario-bound simulations (Shimako et al., 2017). The DLCIA method developed by Levasseur et al. 626 (Levasseur et al., 2010) is currently one of the most recognised and sophisticated approaches, featuring period-627 specific CFs. In addition, calendar-specifications can be relevant to assess and compare the evolution of impacts and/or background concentrations over time (e.g. 1990 Kyoto Protocol and the 1750 IPCC reference years for 628 629 climate change).

630 Time-dependent characterisation functions

631 Recent works (Shimako et al., 2017; Shimako et al., 2018; Shimako et al., 2016) have proposed to come back to 632 the origins of impact simulation tools and adapt them by adding temporal information in the LCIA phase. The 633 idea is to consider the opportunities of using DLCIs as inputs for DLCIA models. Such a DLCIA model has been 634 proposed to assess toxicity impacts (human and ecotoxicity) by Shimako et al. (2017) and has been applied in a 635 full DLCA study. The model reintroduces the time dimension for fate modelling of substances in the 636 environment, providing the temporal distributions of substances in different environmental compartments. The 637 physical parameters for the calculation of fate, exposure and effect factors were taken from the USEtox model. 638 This method doesn't propose period-specific CFs, but directly calculates the impacts by coupling the impact 639 model with all the available information in DLCIs.

640 The definition of ecotoxicity according to time also allows to evaluating the intensity of the impact for different 641 periods of occurrence, which supports the identification of critical periods for potential impacts. The cumulated toxicity then represents the total damage generated over a TH. When compared with conventional USEtox 642 643 results, obtained in steady state conditions, the DLCA results are systematically lower, but toxicity tends towards 644 the conventional results for an infinite TH. Non-persistent substances (generally organic) generate almost all 645 their hazard potential during their periods of emission and disappear more or less rapidly due to the degradation 646 or transfer to sink compartments (removal). In contrast, persistent substances accumulate in environmental 647 compartments during the emission periods and their toxicity potentials remain high after the emissions stop, 648 potentially affecting many human generations.

649 3.4.9. Approach and tools

As was explained in subsection 3.3.2, some examples of using combined DLCI and DLCIA methods have been

published recently for DyPLCA (Shimako et al., 2017; Shimako et al., 2016) and Temporalis (Cardellini et al.,

- 652 2018) respectively for the toxicity and climate change categories. Still, this type of combination is rare and can
- only be done for few impact assessment methods with period-specific characterisation factors or time-dependent
- 654 characterisation functions. Further developments are definitely required here to allow for a comprehensive
- 655 consideration of the dynamics of impacts in future DLCA studies.

656 4. PROPOSED DEVELOPMENT PATHWAYS

It is rather straightforward to define key temporal considerations within the DLCA framework when the challenges of data availability and management are overlooked. Indeed, the general goal can be summarised by a desire to reach the highest level of temporal representativeness and to provide useful information for analysis, when considering the dynamic of systems in all of the model components. It would then seem relevant to:

• Clearly define calendar-based temporal scopes for all flows of a DLCI to outline the periods of elementary flow occurrences that justify the choice for DLCIA methods with specific temporal scopes or THs. This temporal information would also set a clear temporal frame of reference for all stakeholders who want to identify when their decisions will have effects. Moreover, a period of validity for the results of a LCA study should be set as mandatory information to offer an explicit estimation of the period when results can be considered representative and when updates would be necessary.

• Use comprehensive calendar-specific information for the models of the technosphere and ecosphere systems. It would thus be possible to clearly explain when historical data is considered representative. Prospective data based on forecasting strategies and CFs representing future impacts could also be reported explicitly to substantiate the basis for evolution of processes and their temporal scopes. A clear separation between historic and future-related results would then show the proportion of impacts that can only be based on forecasting assumptions.

• Describe the inherent variations of all flows and CFs over a life cycle with the necessary level of detail to 674 minimise the temporal uncertainty of results. Temporal distributions of flows would be defined relative to 675 systems' components for a common framework of assessment, which considers the dynamics of system and 676 impacts that need to be modelled.

Reaching such a comprehensive and complex representation for temporal considerations in the LCA framework would considerably increase our ability to differentiate the impacts of different systems by removing most of the temporal uncertainties from simplifications, but it is probably out of reach and might not be necessary for most comparisons. Consequently, the current challenge lies more in finding the right balance between additional efforts for data collection, modelling complexity and sufficient temporal representativeness. The search for such a "simple but complex enough" implementation strategy is therefore the key to propose the next development steps for temporal considerations in DLCA.

684 4.1. Stepwise approach for temporal considerations with current knowledge

- While many developments can be proposed (see following sub-sections), it is important to recognise that we can already build a strategy from previous ideas and discussions on temporal considerations LCA (section 3). We thus suggest the
- 687following 14 steps and 9 questions within the four standard phases of the LCA framework to help practitioners in the688identification of where and how temporal considerations could be included.
- Figure 3 presents this stepwise general approach, which can be used for any study or system. Sector specific additions have been proposed for some cases like the building sector (Collinge et al., 2013b; Negishi et al., 2018;
- 691 Pittau et al., 2019) and biogenic carbon (Breton et al., 2018; Guest et al., 2013), which could be used in some
- 692 DLCA studies.

693 The colour code is the same as the one used in figure 2 to highlight connections where solid- or white-filled 694 boxes respectively present common and rarer temporal considerations in current LCA studies. Some other 695 remarks are important to use this stepwise approach. First, the chosen technosphere systems in step 1 (S1) is important to identify potential temporal discrepancies and sectors where DLCA is more often useful as explained 696 697 in the introduction (e.g. buildings, energy). Second, the white-filled box of the goal & scope are mostly 698 providing further information on different temporal scopes that are usually not explicitly defined in LCA studies. 699 Third, step 9 (S9) and question 5 (Q5) are the initial places where the need to use a DLCA approach might be 700 identified. Step 12 (S12) and question 7 (Q7) might also highlight such a need. In both cases, different options 701 are available (i.e. S9a, S9b, S12a) depending on the aimed level of detail.

702 The final step (S14) of sensitivity analysis on temporal parameters is certainly useful but currently difficult to 703 implement comprehensively, like what has been proposed by Collet et al. (2014), mainly because there is still a 704 need for deeper investigation of this aspect for all impact categories. Nevertheless, some analyses on 705 technological parameters of the technosphere models are possible and have been carried out for buildings 706 (Asdrubali et al., 2020; Hu, 2018), photovoltaic installations (Louwen et al., 2016) and other renewable energy 707 sources (Pehnt, 2006). A more complete analysis of ecoinvent v2.2 also showed the important variations of GWP 708 when a DLCA was conducted for processes related to wood, biofuels, infrastructure and electricity (Pinsonnault 709 et al., 2014). These examples show that potential technological improvements and increased lifetimes should be 710 investigated in many DLCA studies, but it is not yet possible to provide a full overview of relevant temporal 711 parameters in models.



Figure 3: Stepwise approach to identify where and how to include temporal considerations in the LCA framework. S: Steps / Q: Questions / Green = Yes / Red = No

712

714 4.2. Temporal considerations in the goal and scope definition

715 Temporal considerations, presented in section 3.1, mostly offer partial, implicit and qualitative information about 716 when LCA studies are temporally representative or for when potential impacts are occurring. Temporal scopes of 717 results in LCA studies are sometimes more explicitly defined, but they are not commonly provided, which 718 hinders transparent and fair comparisons among results of different studies (Caffrey and Veal, 2013; Dandres et 719 al., 2012; Huang et al., 2012; Woo et al., 2015). Lack of consistency in the vocabulary that describes the models' 720 components and the linked LCIs or LCIA methods also brings some issues to simplify the exchange of temporal 721 information. These obstacles should be addressed to access the wealth of information and metadata that is 722 currently provided in LCA databases and studies. Two propositions are thus made for potential development 723 pathways:

Recognise and use a common structure and vocabulary to discuss and exchange on the subject of temporal
 considerations in the DLCA framework, databases and studies (see section 2 for propositions).

2. Employ common metadata formats to automate the exchange of temporal information and thus provide access to the wealth of temporal information that is currently provided in LCA databases and studies, as well as to manage the expected significant increase in data requirements for this subject.

A specific example for automation is the development of guidelines to define the different temporal scopes consistently and periods of validity that should be provided in LCA databases for all datasets and studies for all processes. The authors are well aware of the challenge in asking a community to accept a common framework for such a broad subject, but data providers would benefit from the identification of common patterns and of "translation" options between data format.

734 4.3. Time dependent modelling of human activities

Strategies to account for inherent variations and future evolution of systems and impacts have always been implicitly considered in LCA. The mere goal of summing elementary flows over the full life cycle is a testament of this. Nevertheless, most of the current studies show an implicit assumption that human activities and associated elementary flows will not change significantly over their temporal scopes or that such changes do not have to be considered to differentiate the environmental impacts of two products with equivalent functions.

Alternatively, DLCA studies start from the assumption that inherent variations, periods of occurrence and evolution need to be accounted. The basic principle is to consider such levels of temporal considerations with process differentiation, which turns out to be challenging due to the large amount of temporal information 743 needed whenever a comprehensive and detailed description of the life cycle is expected. The temporal 744 differentiation of flows with process-relative temporal distributions has also been shown to be feasible, but has 745 not yet been implemented in commercial databases. Given the current challenges and options, the next steps of 746 development for time-dependent system modelling are suggested, as follows:

747 1. Carry out a comprehensive review of methodologies and approaches where dynamic modelling is 748 considered in other fields of research to identify strategies that might not yet be proposed for the DLCA 749 framework. For instance, DLCA is intrinsically rooted on modelling the dynamics of systems. Many models' 750 components describe a large system, featuring several thousands of processes in the technology matrix and 751 many hundreds of elementary flows in the intervention matrix. The introduction of timed variables in the 752 matrixes and vectors of calculation can induce non-linear trends in the causal relationships. Delays might 753 appear in the datasets (e.g. storage processes) or in the interventions (buffer zones at technosphere/ecosphere 754 interface). The discontinuities form due to temporal switch between technical flows (e.g. seasonal supply) or 755 abrupt release could also arise. All these aspects cause a real issue for solving, simulating and providing 756 DLCA results under a reasonable computation time. Nonetheless, system dynamics is a well-studied topic in 757 applied mathematics and control theory. The introduction of temporal considerations into the field of LCA would thus benefit from the knowledge of these research areas or disciplines. 758

759 2. Provide more process-relative temporal distributions to describe all flows and use these distributions within 760 new computational tools. The identification of key sources for temporal variability within systems is 761 probably the best way to start this work and will increase our knowledge on this subject in an iterative 762 manner. Furthermore, process-relative descriptions should be combined with calendar-specific processes that change automatically when they are no longer representative of the operating technology or activity 763 764 over the considered life cycle (i.e. period of validity). Furthermore, the temporal resolution that is sufficient 765 for such distributions should be balanced with the efforts to describe the models (i.e. data management and 766 gathering).

3. Consider that some technosphere flows or processes might have fixed historical settings when human activities are represented. For example, all elementary flows that are linked to the construction phase of hydro power plants in a country will not have different periods of occurrence if they are linked to past or future products.

4. Identify and define the temporal correlation of flows in current databases. From a mechanistic point of view,
 these relationships exist (e.g. carbon content in CO₂ from tailpipe emission depends on fuel consumption)

and LCA practitioners can use them when creating datasets. By making these relationships explicit, one
 could simplify the introduction of temporal considerations in datasets, as some are intrinsically linked over
 time (e.g. nitrate emissions at the crop level are strongly related to the crop production cycle).

776 Find solutions for temporal considerations with co-product management and allocation. Indeed, the avoided 5. product approach raises the question of how avoided product(s) can be modelled in time. Should it be 777 simultaneous to the co-product or following the co-product creation, assuming that the replacement will take 778 779 place afterwards? A non-physical allocation raises other questions about temporal considerations. For instance, to ensure carbon balance, corrections are made when multi-output processes are split into several 780 781 single-output processes. Artificial positive and negative CO₂ emissions are added up to match the carbon fixations to the carbon content of a product (Weidema, 2018). This approach is, for example, used in the 782 783 econvent database under "At Point of Substitution (APOS)" and "Cut-off" system models (Wernet et al., 784 2016). These allocation options question whether to maintain these flows in DLCIs, and if so, how to position these artificial flows over time. Therefore the period of occurrence will be difficult to justify in 785 786 DLCA.

6. Offer more explicit and complete list of choices made for prospective (or retrospective) modelling and the use of scenarios. The reason for using such modelling approach is to provide results with future (or historic) perspectives that fit more with the objectives of LCA practitioners. It is important to recognise that it is currently challenging to find a consensus on a "best" option for any case study. In such a context, a more achievable goal is to improve the transparency of modelling choices. It would also be useful to separate the elementary flows that are linked to past and present processes from the ones that are based on prospective models. This would clarify the share of impacts issued from modelling assumptions in prospective models.

794 4.4. Inventory calculation: keeping time in the LCI

The recently developed conceptual frameworks and tools (see subsections 3.3.1 and 3.3.2) employ a common computational structure based on graph search algorithms to calculate DLCIs. This structure uses processrelative temporal distributions to describe the flows within system models. Such a consensus suggests that the computational structure for DLCA and the corresponding tools could become a standard, but implementation challenges are still limiting their use. It thus seems relevant to:

Carry out more DLCA studies with these tools to increase the understanding of the LCA community and to
 develop the use of process-relative descriptions in LCA databases.

802 Check the importance of temporal resolution for flows in DLCI. A LCA system can represent many 2. 803 dynamics, because of the size of the system and the inherent temporal variations of the production 804 processes, emissions and resource consumption, as well as of the environmental mechanisms. This issue has 805 already been identified and discussed in some LCA studies where process dynamics are relevant. For 806 instance, Collet et al. (2014) discussed the necessary match between the emission dynamic and the impact 807 category to justify such temporal considerations. Shimako et al. (2018) dealt with the time step of 808 simulations regarding the impact features and showed the gap between examples of climate change (year) 809 and ecotoxicity (day). Urban traffic is another example of the time-resolution aspect that shows the 810 relevance of intraday dynamic for commuters since they mainly travel at the beginning and the end of the 811 working period. Moreover, let's consider, for the sake of clarity, that both carbon dioxide and particulate 812 matter have an intraday emission dynamic. If this resolution seems suitable for the fate of particulate matter, 813 it is clearly too short for climate change mechanisms, where a yearly resolution would be sufficient. The transportation activity also needs infrastructure, which is defined over decades, adding an even slower 814 815 dynamic to the system. Consequently, urban traffic is a good example of a system that merges multiple time 816 resolutions with fast and slow environmental effects. Investigating different systems with varying timescales 817 will thus be relevant to identify temporal consistency in systems.

818 4.5. Dynamics of impact assessment

Temporal considerations in methods for impact characterisation can be introduced with the choice of specific THs. The recent developments in DLCIA methods have focused on the impact categories of climate change, toxicity and ozone depletion, but there is the need to further explore temporal considerations in the phase of impact assessment for the following subjects:

Identify methods to consistently consider THs in DLCA studies for impact categories where it is relevant. A
 clear definition of the temporal scope covered by the LCIA methods would indeed be useful when impacts
 have strong time dependency. The choice of a TH should be based on the limits that are set in the goal and
 scope of a case study. However, to reduce value-laden choices, sensitivity analysis should be encouraged to
 assess the temporal variability in results. For instance, by determining the use of different THs or setting
 different end-years in the dynamic results when using period specific CFs.

Propose a clear list of the relevant time scales for each LCIA category to fix database requirements in the
 definition of elementary flows for any datasets. As explained before, environmental mechanisms for
 different impacts of substances will occur within diverse temporal scopes. These specific periods for each

- impact category can therefore provide guidance on the required resolution of temporal distributions todescribe the elementary flows of LCIs, while minimising the temporal uncertainty.
- 3. Update the considered background concentrations in ecosphere models (i.e. impact assessment methods) when they substantially change the obtained CFs for an impact category. Sensitivity analyses could be performed on past and current concentration levels in order to assess temporal variability of CFs, and to propose, if necessary, updated values for prospective and/or retrospective studies.
- 4. Propose strategies for transparent use of prospective assumptions in ecosphere models. Identifying the parameters that were or will be affected by historic or future modifications of the environment could be particularly relevant in the context of forecasting system evolution. Temporal parameters may be based on, for example, projections of population density, or scenario-bound background concentrations. A clear identification and transparent disclosure of the temporal parameters that most affect the calculation of CFs could indeed be an important added value for impact assessment methods.
- Collaboration between experts of LCA databases, LCI computation and LCIA methods should be strengthened
 to develop a common framework for temporal considerations in any impact assessment methods.
- 846 4.6. Summary of potential development paths for temporal considerations in DLCA
- Table 4 presents a summary of the proposed developments from sections 4.2 to 4.5 with their main purposes
- along the different phases of the LCA framework and a qualitative assessment of the expected level of challenge
- to reach these targets. This assessment goes from + (i.e. basic efforts) to +++ (significant efforts).

851	Table 4: Summary	of proposed	development	paths for tempora	l consideration in	a DLCA framework
-----	------------------	-------------	-------------	-------------------	--------------------	------------------

Pro	posed development paths				
		Defining the	Considering	Increasing the	Challenge
		temporal	the dynamics	temporal	level
		scope	of systems	representativeness	
4.2	Time in the goal and scope definition				
1.	Use of a standard glossary to describe temporal	v	v		++
	considerations in DLCA databases and studies	Λ	Λ		TT
2.	Use of common metadata descriptions to	v	v		++
_	automate the exchange of temporal information	Α	Λ		
4.3	Time dependent modelling of human activities				
1.	Investigate how other fields of research are		x		+
	modelling the dynamic of systems		1		·
2.	Provide more process-relative temporal		x		++
	distributions in DLCA studies to describe flows		1		
3.	Identify the fixed historical nature of some	х	х	х	++
	technosphere processes				
4.	Describe the temporal correlation of flows in		х		++
	datasets of LCA databases				
5.	Find solution(s) for allocation that can be		Х	Х	+++
	accepted by the LCA community				
6.	Offer more explicit and complete list of choices	Х	Х	Х	+++
	made when prospective modelling is used				
4.4	Inventory calculation with temporal properties				
1.	Carry out more DLCA studies with current		Х	Х	++
	approach and tool to increase understanding				
2.	Evaluate the importance of temporal resolution		Х		+++
	in the description of DLCIs for DLCIA				
4.5	Dynamics of impact assessment				
1.	Consistent use of THs in DLCA studies with	Х		Х	+++
	sensitivity assessment				
2.	Provide lists of relevant time scales for each	Х			++
2	impact category				
3.	Update CFs when changes in background			Х	+
4	concentration have substantial effects				
4.	Other more explicit and complete list of choices	Х	Х	Х	+++
	made when prospective modelling is used				

852

853 5. CONCLUSIONS

854 Considerable efforts have been made in the last 20 years to include temporal considerations within the LCA 855 framework and to show that accounting for such aspects significantly affects the results of, at least, some case 856 studies. For instance, LCA studies on systems with long lifespan (e.g. buildings) can benefit from models and 857 tools where the dynamics of energy flows are considered with more details (i.e. variations and evolution). 858 Periods of validity for datasets, which represent rapidly progressing technologies (e.g. photovoltaic cells), are 859 important temporal information, provided in some LCA databases. Furthermore, dynamic LCIA methods have 860 now been developed to account for impacts that vary significantly when the timing of emission change. Overall, 861 the suggested approaches, tools and strategies increase the temporal representativeness of LCA studies and decrease the temporal uncertainty of models the technosphere and its impacts. Nevertheless, their uses in current LCA studies are still uncommon, which can be explained mainly by a lack of consistent descriptions and the challenges of gathering temporal information.

865 With that in mind, we offer some propositions for the next steps of developments in the DLCA framework. A glossary is proposed to build a common and consistent understanding on the key concepts that often come up in 866 867 discussions on the subject. This common understanding should then help in the use of the already available 868 information that can be found in LCA databases and studies under different names. The consistent description of 869 this metadata should also simplify the automated exchange of information between different software options 870 and practitioners. The temporal boundaries of DLCIs (i.e. temporal scope) should be defined within a calendar-871 based description (e.g. between 1990 and 2020) to improve the potential for representativeness of the impact 872 assessments and the fairness of comparison between systems. In addition, our overview on temporal 873 considerations in the LCI phase suggests that a preferred pathway seems to emerge in the computational 874 approach (i.e. graph search algorithms) for DLCA, but it will require the use of process-relative temporal distributions to describe flows in datasets (i.e. input format). This information should then provide temporal 875 876 distributions for all elementary flows. A balance between necessary data collection efforts and reduction of 877 uncertainties should define the temporal resolution of such distributions. It will also be important to consider the 878 chosen DLCIA methods when selecting the temporal resolutions of flows. Lastly, the current development of the 879 DLCIA methods should continue by pursuing the estimation of uncertainty and variability that comes up in all 880 impact categories when temporal information is not provided to describe the input LCI. It is then recommended 881 to identify a relevant level of temporal resolution that would minimise the temporal uncertainty of the models for 882 impact assessments.

883 6. ACKNOWLEDGEMENTS

The authors want to thank both the Empa and IFPEN institutes for funding their work, which allowed time to work on the bulk of this review paper. The first author would also like to highlight that some of this work stems from his time at MINES ParisTech when he was under the supervision of Isabelle Blanc.

888 7. REFERENCES

- Albers A, Avadí A, Benoist A, Collet P, Hélias A. Modelling dynamic soil organic carbon flows of
 annual and perennial energy crops to inform energy-transport policy scenarios in
 France. Science of The Total Environment 2020; 718: 135278.
- Albers A, Collet P, Benoist A, Hélias A. Back to the future: dynamic full carbon accounting applied
 to prospective bioenergy scenarios. The International Journal of Life Cycle Assessment
 2019a.
- Albers A, Collet P, Benoist A, Hélias A. Data and non-linear models for the estimation of biomass
 growth and carbon fixation in managed forests. Data in Brief 2019b; 23: 103841.
- Albers A, Collet P, Lorne D, Benoist A, Hélias A. Coupling partial-equilibrium and dynamic
 biogenic carbon models to assess future transport scenarios in France. Applied Energy
 2019c; 239: 316-330.
- Alvarez-Gaitan JP, Short MD, Peters GM, MacGill I, Moore S. Consequential cradle-to-gate carbon
 footprint of water treatment chemicals using simple and complex marginal technologies
 for electricity supply. The International Journal of Life Cycle Assessment 2014; 19: 1974 1984.
- Amor MB, Gaudreault C, Pineau P-O, Samson R. Implications of integrating electricity supply
 dynamics into life cycle assessment: a case study of renewable distributed generation,
 2014.
- Anand CK, Amor B. Recent developments, future challenges and new research directions in LCA
 of buildings: A critical review. Renewable & Sustainable Energy Reviews 2017; 67: 408 416.
- Arbault D, Riviere M, Rugani B, Benetto E, Tiruta-Barna L. Integrated earth system dynamic
 modeling for life cycle impact assessment of ecosystem services. Science of the Total
 Environment 2014; 472: 262-272.
- Arodudu O, Helming K, Wiggering H, Voinov A. Towards a more holistic sustainability
 assessment framework for agro-bioenergy systems A review. Environmental Impact
 Assessment Review 2017; 62: 61-75.
- Asdrubali F, Baggio P, Prada A, Grazieschi G, Guattari C. Dynamic life cycle assessment modelling
 of a NZEB building. Energy 2020; 191: 116489.
- AzariJafari H, Yahia A, Ben Amor M. Life cycle assessment of pavements: reviewing research
 challenges and opportunities. Journal of Cleaner Production 2016; 112: 2187-2197.
- Bakas I, Hauschild MZ, Astrup TF, Rosenbaum RK. Preparing the ground for an operational handling of long-term emissions in LCA. International Journal of Life Cycle Assessment 2015; 20: 1444-1455.
- Bauer C, Hofer J, Althaus H-J, Del Duce A, Simons A. The environmental performance of current and future passenger vehicles: Life cycle assessment based on a novel scenario analysis framework. Applied Energy 2015; 157: 871-883.
- Beames A, Broekx S, Heijungs R, Lookman R, Boonen K, Van Geert Y, et al. Accounting for land use efficiency and temporal variations between brownfield remediation alternatives in
 life-cycle assessment. Journal of Cleaner Production 2015; 101: 109-117.
- Beloin-Saint-Pierre D, Heijungs R, Blanc I. The ESPA (Enhanced Structural Path Analysis)
 method: a solution to an implementation challenge for dynamic life cycle assessment
 studies. The International Journal of Life Cycle Assessment 2014; 19: 861-871.
- Beloin-Saint-Pierre D, Padey P, Périsset B, Medici V. Considering the dynamics of electricity
 demand and production for the environmental benchmark of Swiss residential buildings
 that exclusively use electricity. IOP Conference Series: Earth and Environmental Science
 2019; 323: 012096.
- Beloin-Saint-Pierre D, Levasseur A, Margni M, Blanc I. Implementing a Dynamic Life Cycle
 Assessment Methodology with a Case Study on Domestic Hot Water Production. Journal
 of Industrial Ecology 2017.
- Benoist A. Adapting Life-Cycle Assessment to biofuels: some elements from the first generation
 case. École Nationale Supérieure des Mines de Paris, 2009.

- Bessou C, Basset-Mens C, Tran T, Benoist A. LCA applied to perennial cropping systems: a review
 focused on the farm stage. International Journal of Life Cycle Assessment 2013; 18: 340 361.
- Bessou C, Ferchaud F, Gabrielle B, Mary B. Biofuels, greenhouse gases and climate change. A
 review. Agronomy for Sustainable Development 2011; 31: 1-79.
- Boulay A-M, Bayart J-B, Bulle C, Franceschini H, Motoshita M, Muñoz I, et al. Analysis of water use
 impact assessment methods (part B): applicability for water footprinting and decision
 making with a laundry case study. The International Journal of Life Cycle Assessment
 2015; 20: 865-879.
- Brandao M, Levasseur A. Assessing Temporary Carbon Storage in Life Cycle Assessment and
 Carbon Footprinting: Outcomes of an expert workshop. Luxembourg: Joint Research
 Centre Institute for Environment and Sustainability, 2011.
- Breton C, Blanchet P, Amor B, Beauregard R, Chang W-S. Assessing the Climate Change Impacts
 of Biogenic Carbon in Buildings: A Critical Review of Two Main Dynamic Approaches.
 Sustainability 2018; 10.
- Bright RM, Cherubini F, Stromman AH. Climate impacts of bioenergy: Inclusion of carbon cycle
 and albedo dynamics in life cycle impact assessment. Environmental Impact Assessment
 Review 2012; 37: 2-11.
- Bright RM, Strømman AH, Peters GP. Radiative Forcing Impacts of Boreal Forest Biofuels: A
 Scenario Study for Norway in Light of Albedo. Environmental Science & Technology
 2011; 45: 7570-7580.
- BSI. Specification for the assessment of the life cycle greenhouse gas emissions of goods and
 services. PAS 2050:2011, London, 2011.
- Caffrey KR, Veal MW. Conducting an Agricultural Life Cycle Assessment: Challenges and
 Perspectives. Scientific World Journal 2013.
- Canals LMI, Bauer C, Depestele J, Dubreuil A, Knuchel RF, Gaillard G, et al. Key elements in a
 framework for land use impact assessment within LCA. International Journal of Life
 Cycle Assessment 2007; 12: 5-15.
- Cardellini G, Mutel CL, Vial E, Muys B. Temporalis, a generic method and tool for dynamic Life
 Cycle Assessment. Science of The Total Environment 2018; 645: 585-595.
- Cheah LW, Ieee. Materials Flow Analysis and Dynamic Life-cycle Assessment of Lightweight
 Automotive Materials in the US Passenger Vehicle Fleet, 2009.
- Cherubini F, Guest G, Stromman AH. Application of probability distributions to the modeling of
 biogenic CO2 fluxes in life cycle assessment. Global Change Biology Bioenergy 2012; 4:
 784-798.
- Cherubini F, Peters GP, Berntsen T, StrØMman AH, Hertwich E. CO2 emissions from biomass
 combustion for bioenergy: atmospheric decay and contribution to global warming. GCB
 Bioenergy 2011; 3: 413-426.
- Choi JK, Friley P, Alfstad T. Implications of energy policy on a product system's dynamic life cycle environmental impact: Survey and model. Renewable & Sustainable Energy
 Reviews 2012; 16: 4744-4752.
- Chowdhury RB, Moore GA, Weatherley AJ, Arora M. Key sustainability challenges for the global
 phosphorus resource, their implications for global food security, and options for
 mitigation. Journal of Cleaner Production 2017; 140: 945-963.
- Collet P. Analyse de Cycle de Vie de la valorisation énergétique de la biomasse algale : prise en compte des aspects dynamiques dans l'étape d'inventaire, 2012.
- Collet P, Hélias A, Lardon L, Steyer J-P. Time and Life Cycle Assessment: How to Take Time into
 Account in the Inventory Step? In: Finkbeiner M, editor. Towards Life Cycle
 Sustainability Management. Springer Netherlands, Dordrecht, 2011, pp. 119-130.
- Collet P, Lardon L, Steyer JP, Helias A. How to take time into account in the inventory step: a
 selective introduction based on sensitivity analysis. International Journal of Life Cycle
 Assessment 2014; 19: 320-330.
- Collinge W, Landis AE, Jones AK, Schaefer LA, Bilec MM. Indoor environmental quality in a
 dynamic life cycle assessment framework for whole buildings: Focus on human health
 chemical impacts. Building and Environment 2013a; 62: 182-190.

- Collinge WO, DeBois JC, Sweriduk ME, Landis AE, Jones AK, Schaefer LA, et al. Measuring Whole Building Performance with Dynamic LCA: a Case Study of a Green University Building. In:
 Ventura A, de la Roche C, editors. International Symposium on Life Cycle Assessment and
 Construction. RILEM, 2012, pp. 309-317.
- Collinge WO, Landis AE, Jones AK, Schaefer LA, Bilec MM. Dynamic life cycle assessment:
 framework and application to an institutional building. International Journal of Life Cycle
 Assessment 2013b; 18: 538-552.
- Collinge WO, Rickenbacker HJ, Landis AE, Thiel CL, Bilec MM. Dynamic Life Cycle Assessments of
 a Conventional Green Building and a Net Zero Energy Building: Exploration of Static,
 Dynamic, Attributional, and Consequential Electricity Grid Models. Environmental
 Science & Technology 2018; 52: 11429-11438.
- 1007Daly HE, Scott K, Strachan N, Barrett J. Indirect CO2 Emission Implications of Energy System1008Pathways: Linking IO and TIMES Models for the UK. Environmental Science &1009Technology 2015; 49: 10701-10709.
- Dandres T, Gaudreault C, Tirado-Seco P, Samson R. Assessing non-marginal variations with
 consequential LCA: Application to European energy sector. Renewable and Sustainable
 Energy Reviews 2011; 15: 3121-3132.
- Dandres T, Gaudreault C, Tirado-Seco P, Samson R. Macroanalysis of the economic and environmental impacts of a 2005-2025 European Union bioenergy policy using the GTAP model and life cycle assessment. Renewable & Sustainable Energy Reviews 2012; 16: 1180-1192.
- de Faria ABB, Sperandio M, Ahmadi A, Tiruta-Bama L. Evaluation of new alternatives in
 wastewater treatment plants based on dynamic modelling and life cycle assessment
 (DM-LCA). Water Research 2015; 84: 99-111.
- Dyckhoff H, Kasah T. Time Horizon and Dominance in Dynamic Life Cycle Assessment. Journal of
 Industrial Ecology 2014; 18: 799-808.
- Earles JM, Halog A, Ince P, Skog K. Integrated Economic Equilibrium and Life Cycle Assessment
 Modeling for Policy-based Consequential LCA. Journal of Industrial Ecology 2013; 17:
 375-384.
- Elshout PMF, Zelm Rv, Balkovic J, Obersteiner M, Schmid E, Skalsky R, et al. Greenhouse-gas
 payback times for crop-based biofuels. Nature Climate Change 2015; 5: 604-610.
- Espinosa N, Hösel M, Angmo D, Krebs FC. Solar cells with one-day energy payback for the factories of the future. Energy & Environmental Science 2012; 5: 5117-5132.
- Fantke P, Aurisano N, Bare J, Backhaus T, Bulle C, Chapman PM, et al. Toward harmonizing
 ecotoxicity characterisation in life cycle impact assessment. Environmental Toxicology
 and Chemistry 2018; 37: 2955-2971.
- Fantke P, Jolliet O. Life cycle human health impacts of 875 pesticides. International Journal of
 Life Cycle Assessment 2016; 21: 722-733.
- Fantke P, Jolliet O, Apte JS, Hodas N, Evans J, Weschler CJ, et al. Characterizing Aggregated
 Exposure to Primary Particulate Matter: Recommended Intake Fractions for Indoor and
 Outdoor Sources. Environmental Science & Technology 2017; 51: 9089-9100.
- Fantke P, Juraske R, Antón A, Friedrich R, Jolliet O. Dynamic Multicrop Model to Characterize
 Impacts of Pesticides in Food. Environmental Science & Technology 2011; 45: 88428849.
- Fantke P, McKone TE, Tainio M, Jolliet O, Apte JS, Stylianou KS, et al. Global Effect Factors for
 Exposure to Fine Particulate Matter. Environmental Science & Technology 2019; 53:
 6855-6868.
- Fantke P, Wieland P, Juraske R, Shaddick G, Itoiz ES, Friedrich R, et al. Parameterization Models
 for Pesticide Exposure via Crop Consumption. Environmental Science & Technology
 2012; 46: 12864-12872.
- Fantke P, Wieland P, Wannaz C, Friedrich R, Jolliet O. Dynamics of pesticide uptake into plants:
 From system functioning to parsimonious modeling. Environmental Modelling &
 Software 2013; 40: 316-324.
- Fearnside PM. Time preference in global warming calculations: a proposal for a unified index.
 Ecological Economics 2002; 41: 21-31.

- Fernandez-Mena H, Nesme T, Pellerin S. Towards an Agro-Industrial Ecology: A review of nutrient flow modelling and assessment tools in agro-food systems at the local scale.
 Science of the Total Environment 2016; 543: 467-479.
- Finkbeiner M, Ackermann R, Bach V, Berger M, Brankatschk G, Chang Y-J, et al. Challenges in Life
 Cycle Assessment: An Overview of Current Gaps and Research Needs. In: Klöpffer W,
 editor. Background and Future Prospects in Life Cycle Assessment. Springer
 Netherlands, Dordrecht, 2014, pp. 207-258.
- Finnveden G, Hauschild MZ, Ekvall T, Guinee J, Heijungs R, Hellweg S, et al. Recent developments
 in Life Cycle Assessment. Journal of Environmental Management 2009; 91: 1-21.
- Fitzpatrick JJ. Environmental sustainability assessment of using forest wood for heat energy in
 Ireland. Renewable & Sustainable Energy Reviews 2016; 57: 1287-1295.
- 1062Fouquet M, Levasseur A, Margni M, Lebert A, Lasvaux S, Souyri B, et al. Methodological1063challenges and developments in LCA of low energy buildings: Application to biogenic1064carbon and global warming assessment. Building and Environment 2015; 90: 51-59.
- Frijia S, Guhathakurta S, Williams E. Functional Unit, Technological Dynamics, and Scaling
 Properties for the Life Cycle Energy of Residences. Environmental Science & Technology
 2012; 46: 1782-1788.
- Frischknecht R, Jolliet O. Global Guidance for Life Cycle Impact Assessment Indicators. 1, Paris,
 2016.
- 1070Fthenakis V, Alsema E. Photovoltaics energy payback times, greenhouse gas emissions and1071external costs: 2004 early 2005 status. Progress in Photovoltaics 2006; 14: 275-280.
- Garcia R, Gregory J, Freire F. Dynamic fleet-based life-cycle greenhouse gas assessment of the
 introduction of electric vehicles in the Portuguese light-duty fleet. International Journal
 of Life Cycle Assessment 2015; 20: 1287-1299.
- Guest G, Cherubini F, Strømman AH. Global Warming Potential of Carbon Dioxide Emissions
 from Biomass Stored in the Anthroposphere and Used for Bioenergy at End of Life.
 Journal of Industrial Ecology 2013; 17: 20-30.
- Guest G, Stromman AH. Climate Change Impacts Due to Biogenic Carbon: Addressing the Issue of
 Attribution Using Two Metrics With Very Different Outcomes. Journal of Sustainable
 Forestry 2014; 33: 298-326.
- Hauschild M, Goedkoop M, Guinée J, Heijungs R, Huijbregts M, Jolliet O, et al. Identifying best
 existing practice for characterisation modeling in life cycle impact assessment. The
 International Journal of Life Cycle Assessment 2013; 18: 683-697.
- Hauschild MZ. Assessing environmental impacts in a life-cycle perspective. Environmental
 Science & Technology 2005; 39: 81A-88A.
- 1086 Hauschild MZ, Huijbregts MAJ. Life Cycle Impact Assessment: Springer Netherlands, 2015.
- Hauschild MZ, Potting J, Hertel O, Schopp W, Bastrup-Birk A. Spatial differentiation in the
 characterisation of photochemical ozone formation The EDIP2003 methodology.
 International Journal of Life Cycle Assessment 2006; 11: 72-80.
- Heeren N, Jakob M, Martius G, Gross N, Wallbaum H. A component based bottom-up building
 stock model for comprehensive environmental impact assessment and target control.
 Renewable & Sustainable Energy Reviews 2013; 20: 45-56.
- Heijungs R, Suh S. The Computational Structure of Life Cycle Assessment. Vol 11. Dordrecht, The
 Netherlands: Kluwer Academic, 2002.
- Helin T, Sokka L, Soimakallio S, Pingoud K, Pajula T. Approaches for inclusion of forest carbon
 cycle in life cycle assessment a review. Global Change Biology Bioenergy 2013; 5: 475 486.
- Hellweg S, Frischknecht R. Evaluation of Long-Term Impacts in LCA. In: LCA DFo, editor. 22nd
 Discussion Forum on LCA. Discussion Forum on LCA, Zurich, 2004, pp. 339-341.
- Hellweg S, Hofstetter TB, Hungerbuhler K. Discounting and the environment Should current
 impacts be weighted differently than impacts harming future generations? International
 Journal of Life Cycle Assessment 2003a; 8: 8-18.
- Hellweg S, Hofstetter TB, Hungerbuhler K. Discounting and the environment should current
 impacts be weighted differently than impacts harming future generations? The
 International Journal of Life Cycle Assessment 2003b; 8: 8.

- Hellweg S, Hofstetter TB, Hungerbühler K. Time-dependent life-cycle assessment of slag landfills
 with the help of scenario analysis: the example of Cd and Cu. Journal of Cleaner
 Production 2005; 13: 301-320.
- Hellweg S, Milà i Canals L. Emerging approaches, challenges and opportunities in life cycle
 assessment. Science 2014; 344: 1109.
- Herfray G, Peuportier B. Evaluation of Electricity Related Impacts Using a Dynamic LCA Model.
 In: Ventura A, de la Roche C, editors. International Symposium on Life Cycle Assessment and Construction. RILEM, 2012, pp. 300-308.
- Herrchen M. Perspective of the systematic and extended use of temporal and spatial aspects in
 LCA of long-lived products. Chemosphere 1998; 37: 265-270.
- Hertwich EG, Gibon T, Bouman EA, Arvesen A, Suh S, Heath GA, et al. Integrated life-cycle
 assessment of electricity-supply scenarios confirms global environmental benefit of low carbon technologies. 2015.
- Hofstetter P, Baumgartner T, Scholz RW. Modelling the valuesphere and the ecosphere:
 Integrating the decision makers' perspectives into LCA. The International Journal of Life
 Cycle Assessment 2000; 5: 161.
- Holmquist H, Fantke P, Cousins IT, Owsianiak M, Liagkouridis I, Peters GM. An (Eco)Toxicity Life
 Cycle Impact Assessment Framework for Per- And Polyfluoroalkyl Substances.
 Environmental Science & Technology 2020; 54: 6224-6234.
- Hoxha E, Jusselme T, Andersen M, Rey E. Introduction of a dynamic interpretation of building
 LCA results: the case of the smart living (lab) building in Fribourg, Switzerland.
 Sustainable Built Environment (SBE), 2016.
- 1128Hu M. Dynamic life cycle assessment integrating value choice and temporal factors—A case1129study of an elementary school. Energy and Buildings 2018; 158: 1087-1096.
- Huang CL, Vause J, Ma HW, Yu CP. Using material/substance flow analysis to support sustainable
 development assessment: A literature review and outlook. Resources Conservation and
 Recycling 2012; 68: 104-116.
- Huijbregts MAJ, Guinee JB, Reijnders L. Priority assessment of toxic substances in life cycle
 assessment. III: Export of potential impact over time and space. Chemosphere 2001; 44:
 59-65.
- Huijbregts MAJ, Steinmann ZJN, Elshout PMF, Stam G, Verones F, Vieira MDM, et al. ReCiPe2016.
 A harmonized life cycle impact assessment method at midpoint and endpoint level.
 National Institute for Public Health and the Environment, Nijmegen, 2016.
- Igos E, Rugani B, Rege S, Benetto E, Drouet L, Zachary D, et al. Integrated environmental
 assessment of future energy scenarios based on economic equilibrium models.
 Metallurgical Research & Technology 2014; 111: 179-189.
- Igos E, Rugani B, Rege S, Benetto E, Drouet L, Zachary DS. Combination of equilibrium models
 and hybrid life cycle-input-output analysis to predict the environmental impacts of
 energy policy scenarios. Applied Energy 2015; 145: 234-245.
- Inyim P, Pereyra J, Bienvenu M, Mostafavi A. Environmental assessment of pavement
 infrastructure: A systematic review. Journal of Environmental Management 2016; 176:
 128-138.
- IPCC. Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the
 Fifth Assessment Report of the Intergovernmental Panel on Climate Change. In: Press CU,
 editor. IPCC, Cambridge, United Kingdom and New York, NY, USA, 2013, pp. 1535.
- IPCC. Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to the
 Fifth Assessment Report of the Intergovernmental Panel on Climate Change. IPCC,
 Geneva, Switzerland, 2014, pp. 151.
- 1154 ISO14040. Life cycle assessment Principles and framework. Environmental management, 2006.
- ISO14044. Life cycle assessment Requirements and guidelines. Environmental management,
 2006.
- Joint Research Center IfEaS, European Commission. ILCD Handbook General Guide for Life
 Cycle Assessment Detailed Guidance. Vol EUR 24708 EN. Luxembourg: Publication
 office of the european union, 2010.

- 1160Joint Research Center IfEaS, European Commission. ILCD handbook: Recommendations for Life1161Cycle Impact Assessment in the European context. Luxembourg: EC-JRC, 2011.
- 1162Jolliet O, Margni M, Charles R, Humbert S, Payet J, Rebitzer G, et al. IMPACT 2002+: A new life1163cycle impact assessment methodology. International Journal of Life Cycle Assessment11642003; 8: 324-330.
- Jolliet O, Saadé M, Crettaz P, Shaked S. Analyse du cycle de vie Comprendre et réaliser un écobilan. EPFL Lausanne Suisse: Presses polytechniques et universitaires romandes, 2010.
- Jørgensen SV, Hauschild MZ, Nielsen PH. Assessment of urgent impacts of greenhouse gas
 emissions—the climate tipping potential (CTP). The International Journal of Life Cycle
 Assessment 2014; 19: 919-930.
- Jørgensen SV, Hauschild MZ, Nielsen PH. The potential contribution to climate change mitigation
 from temporary carbon storage in biomaterials. The International Journal of Life Cycle
 Assessment 2015; 20: 451-462.
- 1174 Karl AAW, Maslesa E, Birkved M. Environmental performance assessment of the use stage of
 1175 buildings using dynamic high-resolution energy consumption and data on grid
 1176 composition. Building and Environment 2019; 147: 97-107.
- Kendall A. Time-adjusted global warming potentials for LCA and carbon footprints. International
 Journal of Life Cycle Assessment 2012; 17: 1042-1049.
- 1179 Kendall A, Chang B, Sharpe B. Accounting for Time-Dependent Effects in Biofuel Life Cycle
 1180 Greenhouse Gas Emissions Calculations. Environmental Science & Technology 2009; 43:
 1181 7142-7147.
- Kim SJ, Kara S, Hauschild M. Functional unit and product functionality—addressing increase in consumption and demand for functionality in sustainability assessment with LCA. The International Journal of Life Cycle Assessment 2017; 22: 1257-1265.
- 1185 Klöpffer W. Introducing Life Cycle Assessment and its Presentation in 'LCA Compendium'. In:
 1186 Klöpffer W, editor. Background and Future Prospects in Life Cycle Assessment. Springer
 1187 Netherlands, Dordrecht, 2014, pp. 1-37.
- Knapp K, Jester T. Empirical investigation of the energy payback time for photovoltaic modules.
 Solar Energy 2001; 71: 165-172.
- Knoeri C, Waeger PA, Stamp A, Althaus HJ, Weil M. Towards a dynamic assessment of raw
 materials criticality: Linking agent-based demand With material flow supply modelling
 approaches. Science of the Total Environment 2013; 461: 808-812.
- Koellner T, de Baan L, Beck T, Brandao M, Civit B, Margni M, et al. UNEP-SETAC guideline on
 global land use impact assessment on biodiversity and ecosystem services in LCA.
 International Journal of Life Cycle Assessment 2013; 18: 1188-1202.
- Lebailly F, Levasseur A, Samson R, Deschenes L. Development of a dynamic LCA approach for the
 freshwater ecotoxicity impact of metals and application to a case study regarding zinc
 fertilization. International Journal of Life Cycle Assessment 2014; 19: 1745-1754.
- Lenzen M, Dey CJ, Murray SA. Historical accountability and cumulative impacts: the treatment of
 time in corporate sustainability reporting. Ecological Economics 2004; 51: 237-250.
- Levasseur A, Brandao M, Lesage P, Margni M, Pennington D, Clift R, et al. Valuing temporary
 carbon storage. Nature Clim. Change 2012a; 2: 6-8.
- Levasseur A, Cavalett O, Fuglestvedt JS, Gasser T, Johansson DJA, Jørgensen SV, et al. Enhancing
 life cycle impact assessment from climate science: Review of recent findings and
 recommendations for application to LCA. Ecological Indicators 2016; 71: 163-174.
- Levasseur A, Lesage P, Margni M, Brandao M, Samson R. Assessing temporary carbon
 sequestration and storage projects through land use, land-use change and forestry:
 comparison of dynamic life cycle assessment with ton-year approaches. Climatic Change
 2012b; 115: 759-776.
- Levasseur A, Lesage P, Margni M, Deschenes L, Samson R. Considering Time in LCA: Dynamic
 LCA and Its Application to Global Warming Impact Assessments. Environmental Science
 & Technology 2010; 44: 3169-3174.
- Levasseur A, Lesage P, Margni M, Samson R. Biogenic Carbon and Temporary Storage Addressed
 with Dynamic Life Cycle Assessment. Journal of Industrial Ecology 2013; 17: 117-128.

- Liao WJ, van der Werf HMG, Salmon-Monviola J. Improved Environmental Life Cycle Assessment
 of Crop Production at the Catchment Scale via a Process-Based Nitrogen Simulation
 Model. Environmental Science & Technology 2015; 49: 10790-10796.
- Ling-Chin J, Heidrich O, Roskilly AP. Life cycle assessment (LCA) from analysing methodology
 development to introducing an LCA framework for marine photovoltaic (PV) systems.
 Renewable & Sustainable Energy Reviews 2016; 59: 352-378.
- Louwen A, van Sark WGJHM, Faaij APC, Schropp REI. Re-assessment of net energy production
 and greenhouse gas emissions avoidance after 40 years of photovoltaics development.
 Nature Communications 2016; 7: 13728.
- Lueddeckens S, Saling P, Guenther E. Temporal issues in life cycle assessment—a systematic
 review. The International Journal of Life Cycle Assessment 2020.
- 1226Maier M, Mueller M, Yan XY. Introducing a localised spatio-temporal LCI method with wheat1227production as exploratory case study. Journal of Cleaner Production 2017; 140: 492-501.
- Manneh R, Margni M, Deschênes L. Evaluating the relevance of seasonal differentiation of human health intake fractions in life cycle assessment. Integrated Environmental Assessment and Management 2012; 8: 749-759.
- Marvuglia A, Benetto E, Rege S, Jury C. Modelling approaches for consequential life-cycle
 assessment (C-LCA) of bioenergy: Critical review and proposed framework for biogas
 production. Renewable and Sustainable Energy Reviews 2013; 25: 768-781.
- Marvuglia A, Kanevski M, Benetto E. Machine learning for toxicity characterisation of organic
 chemical emissions using USEtox database: Learning the structure of the input space.
 Environment International 2015; 83: 72-85.
- Maurice E, Dandres T, Moghaddam RF, Nguyen K, Lemieux Y, Cherriet M, et al. Modelling of
 Electricity Mix in Temporal Differentiated Life-Cycle-Assessment to Minimize Carbon
 Footprint of a Cloud Computing Service. ICT for Sustainability 2014 (ICT4S-14). Atlantis
 Press, 2014.
- McManus MC, Taylor CM. The changing nature of life cycle assessment. Biomass and Bioenergy 2015; 82: 13-26.
- Mehmeti A, McPhail SJ, Pumiglia D, Carlini M. Life cycle sustainability of solid oxide fuel cells:
 From methodological aspects to system implications. Journal of Power Sources 2016;
 325: 772-785.
- Mendivil R, Fischer U, Hirao M, Hungerbühler K. A New LCA Methodology of Technology
 Evolution (TE-LCA) and its Application to the Production of Ammonia (1950-2000) (8
 pp). The International Journal of Life Cycle Assessment 2006; 11: 98-105.
- Menten F, Tchung-Ming S, Lorne D, Bouvart F. Lessons from the use of a long-term energy model
 for consequential life cycle assessment: The BTL case. Renewable & Sustainable Energy
 Reviews 2015; 43: 942-960.
- Messagie M, Mertens J, Oliveira L, Rangaraju S, Sanfelix J, Coosemans T, et al. The hourly life cycle
 carbon footprint of electricity generation in Belgium, bringing a temporal resolution in
 life cycle assessment. Applied Energy 2014; 134: 469-476.
- Miller SA, Moysey S, Sharp B, Alfaro J. A Stochastic Approach to Model Dynamic Systems in Life
 Cycle Assessment. Journal of Industrial Ecology 2013; 17: 352-362.
- Miotti M, Hofer J, Bauer C. Integrated environmental and economic assessment of current and future fuel cell vehicles. The International Journal of Life Cycle Assessment 2017; 22: 94-110.
- Morais SA, Delerue-Matos C. A perspective on LCA application in site remediation services:
 Critical review of challenges. Journal of Hazardous Materials 2010; 175: 12-22.
- Müller-Wenk R, Brandão M. Climatic impact of land use in LCA—carbon transfers between
 vegetation/soil and air. The International Journal of Life Cycle Assessment 2010; 15:
 172-182.
- Mutel C. Brightway: An open source framework for Life Cycle Assessment. Journal of Open
 Source Software 2017; 2: 236.
- 1267Myhre G, Shindell D, Bréon F-M, Collins W, Fuglestvedt J, Huang J, et al. Anthropogenic and1268Natural Radiative Forcing. In: IPCC, editor. Climate Change 2013: The Physical Science1269Basis. Contribution of Working Group I to the Fifth Assessment Report of the

- Intergovernmental Panel on Climate Change [Stocker, T.F., D. Qin, G.-K. Plattner, M.
 Tignor, S.K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex and P.M. Midgley (eds.)]. IPCC,
 Cambridge, United Kingdom and New York, NY, USA, 2013, pp. 659-740.
- Negishi K, Lebert A, Almeida D, Chevalier J, Tiruta-Barna L. Evaluating climate change pathways
 through a building's lifecycle based on Dynamic Life Cycle Assessment. Building and
 Environment 2019; 164: 106377.
- Negishi K, Tiruta-Barna L, Schiopu N, Lebert A, Chevalier J. An operational methodology for
 applying dynamic Life Cycle Assessment to buildings. Building and Environment 2018;
 144: 611-621.
- Noori M, Tatari O. Development of an agent-based model for regional market penetration
 projections of electric vehicles in the United States. Energy 2016; 96: 215-230.
- Núñez M, Pfister S, Vargas M, Antón A. Spatial and temporal specific characterisation factors for
 water use impact assessment in Spain. The International Journal of Life Cycle
 Assessment 2015; 20: 128-138.
- O'Hare M, Plevin RJ, Martin JI, Jones AD, Kendall A, Hopson E. Proper accounting for time
 increases crop-based biofuels' greenhouse gas deficit versus petroleum. Environmental
 Research Letters 2009; 4: 024001.
- Owens JW. Life-Cycle Assessment in Relation to Risk Assessment: An Evolving Perspective. Risk
 Analysis 1997a; 17: 359-365.
- Owens JW. Life-Cycle Assessment: Constraints on Moving from Inventory to Impact Assessment.
 Journal of Industrial Ecology 1997b; 1: 37-49.
- Owsianiak M, Holm PE, Fantke P, Christiansen KS, Borggaard OK, Hauschild MZ. Assessing
 comparative terrestrial ecotoxicity of Cd, Co, Cu, Ni, Pb, and Zn: The influence of aging
 and emission source. Environmental Pollution 2015; 206: 400-410.
- Pahri SDR, Mohamed AF, Samat A. LCA for open systems: a review of the influence of natural and
 anthropogenic factors on aquaculture systems. International Journal of Life Cycle
 Assessment 2015; 20: 1324-1337.
- Pawelzik P, Carus M, Hotchkiss J, Narayan R, Selke S, Wellisch M, et al. Critical aspects in the life
 cycle assessment (LCA) of bio-based materials Reviewing methodologies and deriving
 recommendations. Resources, Conservation and Recycling 2013; 73: 211-228.
- Pehnt M. Assessing future energy and transport systems: The case of fuel cells Part 2:
 Environmental performance. International Journal of Life Cycle Assessment 2003a; 8:
 365-378.
- 1303Pehnt M. Assessing future energy and transport systems: The case of fuel cells Part I:1304Methodological aspects. International Journal of Life Cycle Assessment 2003b; 8: 283-1305289.
- Pehnt M. Dynamic life cycle assessment (LCA) of renewable energy technologies. Renewable
 Energy 2006; 31: 55-71.
- Peña N, Knudsen MT, Fantke P, Antón A, Hermansen JE. Freshwater ecotoxicity assessment of
 pesticide use in crop production: Testing the influence of modeling choices. Journal of
 Cleaner Production 2019; 209: 1332-1341.
- Pesonen H-L, Ekvall T, Fleischer G, Huppes G, Jahn C, Klos ZS, et al. Framework for scenario
 development in LCA. The International Journal of Life Cycle Assessment 2000; 5: 21.
- Pettersen J, Hertwich EG. Critical review: Life-cycle inventory procedures for long-term release
 of metals. Environmental Science & Technology 2008; 42: 4639-4647.
- Pfister S, Bayer P, Koehler A, Hellweg S. Projected water consumption in future global agriculture: Scenarios and related impacts. Science of The Total Environment 2011; 409: 4206-4216.
- Pigné Y, Gutiérrez TN, Gibon T, Schaubroeck T, Popovici E, Shimako AH, et al. A tool to
 operationalize dynamic LCA, including time differentiation on the complete background
 database. The International Journal of Life Cycle Assessment 2019.
- Pinsonnault A, Lesage P, Levasseur A, Samson R. Temporal differentiation of background
 systems in LCA: relevance of adding temporal information in LCI databases. The
 International Journal of Life Cycle Assessment 2014; 19: 1843-1853.

- Pittau F, Habert G, Iannaccone G. A Life-Cycle Approach to Building Energy Retrofitting: Bio Based Technologies for Sustainable Urban Regeneration. IOP Conference Series: Earth
 and Environmental Science 2019; 290: 012057.
- 1327 PRé va. SimaPro Database Manual Methods Library, 2016, pp. 67.
- 1328Querini F, Benetto E. Combining Agent-Based Modeling and Life Cycle Assessment for the1329Evaluation of Mobility Policies. Environmental Science & Technology 2015; 49: 1744-13301751.
- Reap J, Roman F, Duncan S, Bras B. A survey of unresolved problems in life cycle assessment Part 1. The International Journal of Life Cycle Assessment 2008a; 13: 290-300.
- 1333 Reap J, Roman F, Duncan S, Bras B. A survey of unresolved problems in life cycle assessment 1334 Part 2. International Journal of Life Cycle Assessment 2008b; 13: 374-388.
- 1335Rebitzer G, Ekvall T, Frischknecht R, Hunkeler D, Norris G, Rydberg T, et al. Life cycle1336assessment: Part 1: Framework, goal and scope definition, inventory analysis, and1337applications. Environment International 2004; 30: 701-720.
- Recchioni M, Mathieux F, Goralczyk M, Schau EM. ILCD Data Network and ELCD Database:
 current use and further needs for supporting Environmental Footprint and Life Cycle
 Indicator Projects. Joint Research Centre, Ispra, Italy, 2013, pp. 33.
- Risch E, Gasperi J, Gromaire M-C, Chebbo G, Azimi S, Rocher V, et al. Impacts from urban water
 systems on receiving waters How to account for severe wet-weather events in LCA?
 Water Research 2018; 128: 412-423.
- Roder M, Thornley P. Bioenergy as climate change mitigation option within a 2 degrees C target uncertainties and temporal challenges of bioenergy systems. Energy Sustainability and
 Society 2016; 6.
- Rosenbaum RK, Anton A, Bengoa X, Bjørn A, Brain R, Bulle C, et al. The Glasgow consensus on the
 delineation between pesticide emission inventory and impact assessment for LCA. The
 International Journal of Life Cycle Assessment 2015; 20: 765-776.
- Rosenbaum RK, Bachmann TM, Gold LS, Huijbregts MAJ, Jolliet O, Juraske R, et al. USEtox-the
 UNEP-SETAC toxicity model: recommended characterisation factors for human toxicity
 and freshwater ecotoxicity in life cycle impact assessment. International Journal of Life
 Cycle Assessment 2008; 13: 532-546.
- 1354Roux C, Schalbart P, Assoumou E, Peuportier B. Integrating climate change and energy mix1355scenarios in LCA of buildings and districts. Applied Energy 2016a; 184: 619-629.
- Roux C, Schalbart P, Peuportier B. Accounting for temporal variation of electricity production
 and consumption in the LCA of an energy-efficient house. Journal of Cleaner Production
 2016b; 113: 532-540.
- Roux C, Schalbart P, Peuportier B. Development of an electricity system model allowing dynamic
 and marginal approaches in LCA—tested in the French context of space heating in
 buildings. The International Journal of Life Cycle Assessment 2017; 22: 1177-1190.
- Saez de Bikuña K, Hamelin L, Hauschild MZ, Pilegaard K, Ibrom A. A comparison of land use
 change accounting methods: seeking common grounds for key modeling choices in
 biofuel assessments. Journal of Cleaner Production 2018; 177: 52-61.
- Sandberg NH, Brattebø H. Analysis of energy and carbon flows in the future Norwegian dwelling
 stock. Building Research & Information 2012; 40: 123-139.
- Santero N, Loijos A, Akbarian M, Ochsendorf J. Methos, Impacts, and Opportunities in the
 Concrete Pavement Life Cycle. MIT, USA, 2011, pp. 103.
- Scheuer C, Keoleian GA, Reppe P. Life cycle energy and environmental performance of a new university building: modeling challenges and design implications. Energy and Buildings 2003; 35: 1049-1064.
- Shah V, Ries R. A characterisation model with spatial and temporal resolution for life cycle
 impact assessment of photochemical precursors in the United States. The International
 Journal of Life Cycle Assessment 2009; 14: 313-327.
- Shimako AH, Tiruta-Barna L, Ahmadi A. Operational integration of time dependent toxicity
 impact category in dynamic LCA. Science of The Total Environment 2017; 599-600: 806 819.

- Shimako AH, Tiruta-Barna L, Bisinella de Faria AB, Ahmadi A, Spérandio M. Sensitivity analysis
 of temporal parameters in a dynamic LCA framework. Science of The Total Environment
 2018; 624: 1250-1262.
- Shimako AH, Tiruta-Barna L, Pigne Y, Benetto E, Gutierrez TN, Guiraud P, et al. Environmental
 assessment of bioenergy production from microalgae based systems. Journal of Cleaner
 Production 2016; 139: 51-60.
- Shine KP. The global warming potential—the need for an interdisciplinary retrial. Climatic
 Change 2009; 96: 467-472.
- 1386 Simons A, Bauer C. A life-cycle perspective on automotive fuel cells. Applied Energy 2015; 157:
 1387 884-896.
- Sohn J, Kalbar P, Goldstein B, Birkved M. Defining Temporally Dynamic Life Cycle Assessment: A
 Literature Review. Integrated Environmental Assessment and Management 2020; n/a.
- Sohn JL, Kalbar PP, Banta GT, Birkved M. Life-cycle based dynamic assessment of mineral wool
 insulation in a Danish residential building application. Journal of Cleaner Production
 2017a; 142: 3243-3253.
- Sohn JL, Kalbar PP, Birkved M. Life cycle based dynamic assessment coupled with multiple
 criteria decision analysis: A case study of determining an optimal building insulation
 level. Journal of Cleaner Production 2017b; 162: 449-457.
- Standardisation ECf. Assessment of environmental performance of buildings Calculation method.
 Sustainability of construction works. EN 15978, 2009.
- Stasinopoulos P, Compston P, Newell B, Jones HM. A system dynamics approach in LCA to
 account for temporal effects—a consequential energy LCI of car body-in-whites. The
 International Journal of Life Cycle Assessment 2012; 17: 199-207.
- Su S, Li X, Zhu Y, Lin B. Dynamic LCA framework for environmental impact assessment of
 buildings. Energy and Buildings 2017; 149: 310-320.
- Tessum CW, Marshall JD, Hill JD. A Spatially and Temporally Explicit Life Cycle Inventory of Air
 Pollutants from Gasoline and Ethanol in the United States. Environmental Science &
 Technology 2012; 46: 11408-11417.
- Tiruta-Barna L, Pigne Y, Gutierrez TN, Benetto E. Framework and computational tool for the
 consideration of time dependency in Life Cycle Inventory: proof of concept. Journal of
 Cleaner Production 2016; 116: 198-206.
- 1409 Udo de Haes H, Finnveden G, Goedkoop M, Hauschild M, Hertwich E, Hofstetter P, et al. Life-cycle
 1410 impact assessment: striving towards best practice. In: (SETAC) SoETaC, editor, 2002.
- Vazquez-Rowe I, Marvuglia A, Flammang K, Braun C, Leopold U, Benetto E. The use of temporal dynamics for the automatic calculation of land use impacts in LCA using R programming environment. International Journal of Life Cycle Assessment 2014; 19: 500-516.
- Viebahn P, Lechon Y, Trieb F. The potential role of concentrated solar power (CSP) in Africa and
 Europe-A dynamic assessment of technology development, cost development and life
 cycle inventories until 2050. Energy Policy 2011; 39: 4420-4430.
- 1417 Vigon BW, Tolle DA, Cornaby BW, Latham HC, Harrison CL, Boguski TL, et al. LIFE-CYCLE
 1418 ASSESSMENT: INVENTORY GUIDELINES AND PRINCIPLES. In: EPA, editor. EPA,
 1419 Cincinnati, Ohio, US, 1993.
- 1420 Vogtländer JG, van der Velden NM, van der Lugt P. Carbon sequestration in LCA, a proposal for a
 1421 new approach based on the global carbon cycle; cases on wood and on bamboo. The
 1422 International Journal of Life Cycle Assessment 2014; 19: 13-23.
- Vuarnoz D, Cozza S, Jusselme T, Magnin G, Schafer T, Couty P, et al. Integrating hourly life-cycle
 energy and carbon emissions of energy supply in buildings. Sustainable Cities and
 Society 2018; 43: 305-316.
- 1426 Vuarnoz D, Jusselme T. Temporal variations in the primary energy use and greenhouse gas
 1427 emissions of electricity provided by the Swiss grid. Energy 2018; 161: 573-582.
- Walker SB, Fowler M, Ahmadi L. Comparative life cycle assessment of power-to-gas generation
 of hydrogen with a dynamic emissions factor for fuel cell vehicles. Journal of Energy
 Storage 2015; 4: 62-73.

- Walzberg J, Dandres T, Merveille N, Cheriet M, Samson R. Accounting for fluctuating demand in
 the life cycle assessments of residential electricity consumption and demand-side
 management strategies. Journal of Cleaner Production 2019a; 240: 118251.
- Walzberg J, Dandres T, Merveille N, Cheriet M, Samson R. Assessing behavioural change with
 agent-based life cycle assessment: Application to smart homes. Renewable and
 Sustainable Energy Reviews 2019b; 111: 365-376.
- Weidema B, Bauer C, Hischier R, Mutel C, Nemecek T, Reinhard J, et al. Overview and methodology. Data quality guideline for the ecoinvent database version 3. In: ecoinvent, editor. Ecoinvent Report. The ecoinvent Centre, St. Gallen: The ecoinvent Centre, 2012, pp. 166.
- Weidema BP, Schmidt J, Fantke P, Pauliuk S. On the boundary between economy and
 environment in life cycle assessment. The International Journal of Life Cycle Assessment
 2018; 23: 1839-1846.
- Wernet G, Bauer C, Steubing B, Reinhard J, Moreno-Ruiz E, Weidema B. The ecoinvent database
 version 3 (part I): overview and methodology. The International Journal of Life Cycle
 Assessment 2016; 21: 1218-1230.
- Woo C, Chung Y, Chun D, Seo H, Hong S. The static and dynamic environmental efficiency of
 renewable energy: A Malmquist index analysis of OECD countries. Renewable &
 Sustainable Energy Reviews 2015; 47: 367-376.
- Yang J, Chen B. Global warming impact assessment of a crop residue gasification project—A
 dynamic LCA perspective. Applied Energy 2014; 122: 269-279.
- Yang Y, Suh S. Changes in environmental impacts of major crops in the US. Environmental
 Research Letters 2015; 10.
- Yu B, Sun Y, Tian X. Capturing time effect of pavement carbon footprint estimation in the life
 cycle. Journal of Cleaner Production 2018; 171: 877-883.
- Yuan C, Dornfeld D. Embedded Temporal Difference in Life Cycle Assessment: Case Study on VW
 Golf A4 Car. In: IEEE, editor. IEEE International Symposium on Sustainable Systems and
 Technology, Phoenix, Arizona, US, 2009.
- Yuan C, Wang ED, Zhai Q, Yang F. Temporal discounting in life cycle assessment: A critical review
 and theoretical framework. Environmental Impact Assessment Review 2015; 51: 23-31.
- Zaimes GG, Vora N, Chopra SS, Landis AE, Khanna V. Design of Sustainable Biofuel Processes and
 Supply Chains: Challenges and Opportunities. Processes 2015; 3: 634-663.
- 1463ZhaiP, WilliamsED. DynamicHybridLifeCycleAssessmentofEnergyandCarbonof1464MulticrystallineSiliconPhotovoltaicSystems.EnvironmentalScience& Technology14652010;44:7950-7955.
- I466 Zhai Q, Crowley B, Yuan C, Ieee. Temporal Discounting for Life Cycle Assessment: Differences
 between Environmental Discounting and Economic Discounting. 2011 Ieee International
 Symposium on Sustainable Systems and Technology (Issst) 2011.